



# Evaluating the impact of anthropogenic drivers and meteorological factors on air pollutants by explainable machine learning in Shandong Province, China

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## ABSTRACT

Unexpected haze in North China Plain during the COVID-19 lockdown has been regarded as a natural window to explore the meteorological impact on formatting PM<sub>2.5</sub> pollution but with limitations in quantifying weather elements' contributions. In this study, daily data of six air pollutants (including PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, O<sub>3</sub>, and CO) and six meteorological factors (including temperature, pressure, relative humidity (RH), wind speed (WS), wind direction (WD), and precipitation) from 2015 to 2020 across 16 capital cities in Shandong province, China, was used to drive the Machine Learning and the SHapley Additive exPlanation (SHAP) models. By applying these models, contributions from anthropogenic drivers to pollutant reductions and contributions from meteorological factors to the haze event were investigated. Results show that the COVID-19 lockdown measures reduced concentrations of NO<sub>2</sub>, PM<sub>2.5</sub>, PM<sub>10</sub>, CO and SO<sub>2</sub> by −52.1 %, −40.0 %, −45.5 %, −29.4 % and −38.7 % respectively. On average, an 18.9 % increase in O<sub>3</sub> was observed. PM<sub>2.5</sub> pollution was mainly driven by temperature with a SHAP value of 19.7 µg/m<sup>3</sup>, followed by RH (5.8 µg/m<sup>3</sup>), precipitation (0.9 µg/m<sup>3</sup>), WD (0.3 µg/m<sup>3</sup>), pressure (0.1 µg/m<sup>3</sup>) and WS (0.1 µg/m<sup>3</sup>) during the haze period. Relative to the post-haze period, high-pressure systems coupled with lower temperatures and weakened surface winds hindered the dispersion of PM<sub>2.5</sub> whilst higher RH was in favour of PM<sub>2.5</sub> production during the haze period. This study underscores the intricate interplay between emissions, meteorological conditions, and regulatory measures in air pollution, offering critical insights into future air quality management strategies by air pollution prediction.

## 1. Introduction

Air pollution remains a critical global environmental challenge with substantial implications for public health, ecosystem integrity, and the global economic sustainability (Nel, 2005; Shen et al., 2020). According to the World Health Organization (WHO), approximately 7 million premature deaths occur annually as a result of exposure to air pollutants (Kuehn, 2014), with fine particulate matter (PM<sub>2.5</sub>) and ground-level ozone (O<sub>3</sub>) identified as major contributors (Lelieveld et al., 2015;

Shen et al., 2024). These pollutants are strongly associated with increased incident of respiratory and cardiovascular diseases, highlighting the urgent necessity for effective air quality management strategies at both national and global levels (Burnett et al., 2014; Lelieveld et al., 2015).

The outbreak of the COVID-19 pandemic in late 2019 and the subsequent implementation of widespread lockdown restrictions in early 2020 resulted in drastic reductions in industrial operations, vehicular traffic, and other anthropogenic activities. Those abrupt changes led to

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notable declines in nitrogen dioxide (NO<sub>2</sub>) and particulate matter (PM) concentrations (Shen et al., 2022). Numerous studies have confirmed that both variations in emissions and meteorological conditions affect regional air quality significantly (He et al., 2017; Shen et al., 2022). Despite the significant reduction in anthropogenic emissions, several studies also reported that the North China Plain (NCP) still experienced an unexpected haze event during the early 2020 lockdown. This abnormal haze has been attributed to complex atmospheric chemistry processes and special weather conditions (Le et al., 2020; Shen et al., 2024). As a result, the combination of COVID-19 lockdown measures induced emission reductions and the concurrent severe haze event could be regarded as a ‘natural window’ to better disentangle the respective roles of anthropogenic emission reduction and meteorological drivers in air pollution, thereby informing more effective mitigation strategy. To evaluate the contributions of anthropogenic emission reduction and meteorology to the change in air pollutants during the COVID-19 lockdown, Machine Learning (ML) models (Rybarczyk and Zalakeviciute, 2021; Shi et al., 2021; Wang et al., 2020b; Yang et al., 2021) and chemistry transport model (CTM) simulations (Li et al., 2020; Lv et al., 2020; Marlier et al., 2020; Xing et al., 2020) are two effective approaches by designing comparison between different scenarios, such as a ‘business as usual’ (BAU) scenario to the factual scenario. In contrast, the ML model outperforms the CTM model due to the latter’s reliance on outdated emission inventories and limited spatial resolution. Moreover, ML algorithms can also capture the complex nonlinear relationships between predictors and local features’ characteristics (Shen et al., 2022; Xiao et al., 2018). At last, with the exception of the above advantages, ML models can be integrated with explainable AI tools, such as SHapley Additive exPlanations (SHAP) module, which allow for the quantification of feature importance and enhance the physical interpretability of key drivers or major contributors to air pollution (Hou et al., 2022; Men et al., 2023).

Shandong Province, located on the eastern coast of China, is among the most industrialized regions in the country and has experienced persistent air pollution challenges due to its diverse industrial sectors and coal-dominated energy structure (Weng et al., 2022; Zhou et al., 2021). A national assessment of air quality in 168 major cities in China reported that five cities in Shandong, including Zaozhuang, Linyi, Liaocheng, Jinan and Zibo, ranked among the bottom 20 for air quality in 2019 (MEEP, 2019). Moreover, several cities within the province, including Jinan, Zibo, Jining, and Dezhou, are located along a key air pollution transport channel of the NCP region, further exacerbating regional air quality issues in Shandong Province (Zhou et al., 2021). Given these characteristics, Shandong Province could serve as an ideal study domain for investigating the individual roles of anthropogenic emission reduction and meteorology in the local air quality change during the COVID-19 lockdown.

To evaluate the contribution of anthropogenic emission reduction to the air quality change in Shandong Province, a machine learning framework based on the Gradient Boosting Machine (LightGBM) was employed to construct a BAU scenario during the COVID-19 lockdown. Subsequently, the SHAP interpretability module was coupled with the ML model to identify the major meteorological factors contributing to the haze event in this region. Additionally, to better understand how various meteorological factors synergistically to promote haze formation, the impact of atmospheric physical fields on PM<sub>2.5</sub> changes was analyzed as well.

## 2. Materials and methods

### 2.1. Station data of air pollutants and meteorological factors

Shandong province (114°47.5′–122°42.3′E, 34°22.9′–38°24.01′N), located along the eastern coast of China (Fig. S1), was selected as the research domain. Hourly mass concentrations of six air pollutants, including PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, O<sub>3</sub>, and CO, were collected across 16

cities from China National Environmental Monitoring Center (NEMC) (<http://www.cnemc.cn/>). Those hourly measurements were aggregated into daily values from January 1, 2015, to April 30, 2020. Data quality control procedures were implemented in accordance with the methodology described in Shen et al. (2020). The spatial distribution of air quality sampling sites across 16 cities in Shandong Province is shown in Fig. 1. Daily surface meteorological factors (e.g., Temperature: T, °C; Relative Humidity: RH, %; Wind Speed: WS, m/s; Wind Direction: WD, °; Precipitation: Precip, mm; Pressure: Pres, hPa) were also collected for the same period. For the ML model development, the dataset was divided into the training part (January 1, 2015–December 30, 2019) and the test part (January 1, 2020–April 30, 2020). Note that among 123 National Meteorological Stations in Shandong Province, only 16 stations were selected because they are situated closest to the 16 air pollutant monitoring sites. Meteorological data can be accessed from the National Meteorological Science Data Center (<https://data.cma.cn/>).

### 2.2. ERA5 dataset

In addition, hourly surface meteorological reanalysis data from the ERA5 dataset, provided by the European Centre for Medium-Range Weather Forecasts (ECMWF), were used. Those datasets have a spatial resolution of 0.25° × 0.25° and include variables such as 2m surface temperature, dew-point temperature ( $T_d$ , °C), 10m wind, mean sea level pressure (MSLP) and geopotential height, RH, wind, and air temperature at pressure levels from 850 to 1000 hPa. The data were accessed from the Copernicus Climate Data Store (<https://cds.climate.copernicus.eu/>).

### 2.3. Machine learning (ML) model

The LightGBM, an ensemble ML method based on decision trees, was employed to predict the mass concentrations of six air pollutants in BAU in 2020. The LightGBM has demonstrated strong capability in handling complex nonlinear relationships, performing internal feature selection, and efficiently providing feature importance (Ke et al., 2017). Moreover, its ability to capture site-specific features makes it particularly suitable for predicting pollutants in diverse cities, including those lacking emission inventories or located in the desert and plateau regions (Shen et al., 2022). The input dataset of ML model not only comprises six meteorological factors and six air pollutants but also includes four time variables (e.g., Julian Day (JD), Day of Week (DOW), Holidays, and Chinese New Year (CNY)) from 2015 to 2020. To select an optimal ML model, cross-validation by using the time-series split rolling method executed before the implementation of the final predictions. After the step of cross-validation, the performance of the ML model was evaluated based on root-mean-squared error (RMSE) and Pearson correlation coefficient (PCC). Typically, a ML model’s optimal performance is defined by the lowest RMSE and the highest PCC values.

### 2.4. SHapley additive exPlanation (SHAP) method

Measuring the importance of input features for ML model is as vital as attaining high predictive accuracy. To achieve this goal, the SHAP method, based on game theory, is used to quantify the importance of each feature input (Lundberg et al., 2020). SHAP measures the impact of each feature by comparing prediction outcomes with and without the feature (Hou et al., 2022). This approach provides a detailed understanding of each variable’s contribution to the model’s predictions. For a model with  $n$  variables applied to a sample ( $x_i$ ) and its predicted output  $f(x_i)$ , the prediction function can be expressed as follows:

$$f(x_i) = E_0(f, x) + \sum_{m=1}^n E_m(f, x_i) \quad (1)$$

where  $x_i$  represents the input with variable  $m$  in the prediction model  $f$ , producing the SHAP value  $E_m(f, x_i)$ .  $E_0(f, x)$  is the expected value for the prediction model over the entire dataset. In general, the SHAP value

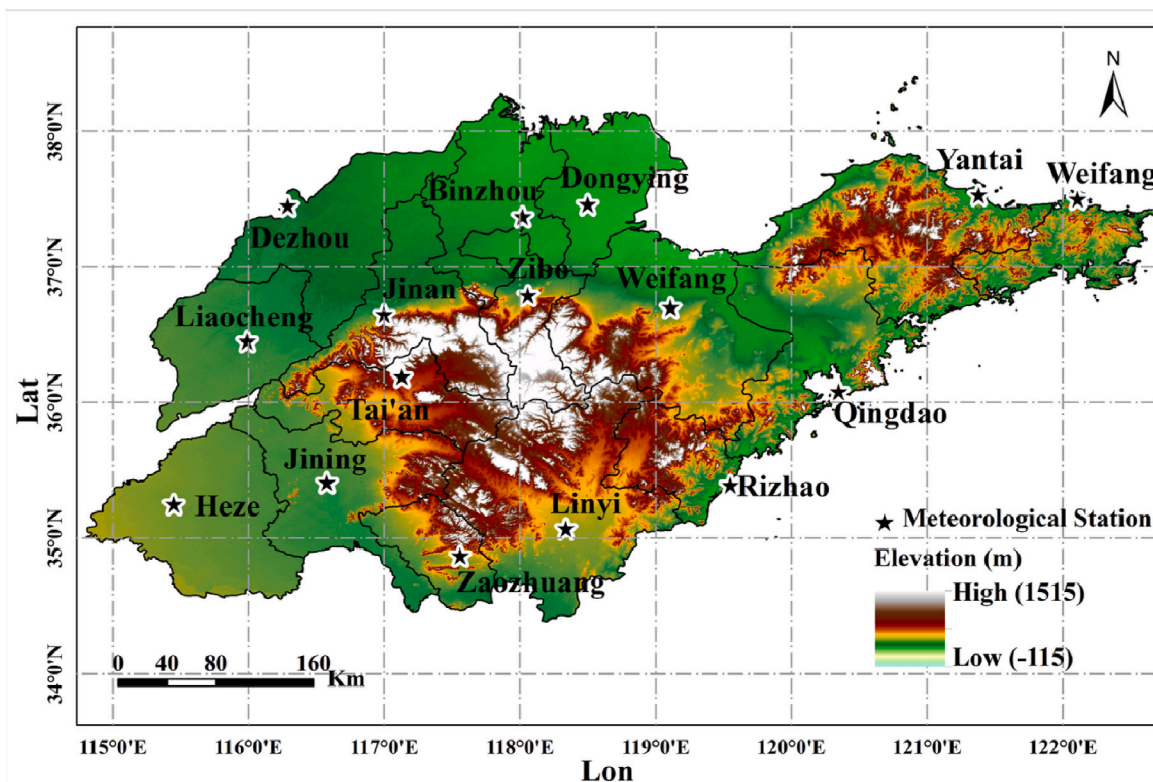


Fig. 1. The locations of 16 surface meteorological observation sites in Shandong Province, China. The color bar indicates the altitude in Shandong.

explains the contribution of each variable to the prediction. This contribution is measured from the base value (the predicted mean of the expected value) to the final model output. Variables that increase the prediction are shown as positive, while those that lower it are displayed as negative.

## 2.5. Specification of the study period

In early 2020, the COVID-19 outbreak prompted local authorities nationwide to implement diverse restrictive measures. These measures resulted in a significant reduction in population movement, industrial production and economic activities (Nie et al., 2021). Notably, on January 25, 2020, Shandong Province initiated a Level I response to address a major public health emergency. Subsequently, on 8 March, the emergency response level for preventing and controlling the Candidatus Pneumonia outbreak in Shandong Province was elevated to Level II. Considering the above two different levels, we defined the entire Level I period as the ‘Total COVID-lockdown’ for analysis in section 3.1. Detailed classifications of the different lockdown phases are shown in Table S1. The CNY period, defined by the lunar calendar, typically starts about a week before and lasts approximately 10 days after the CNY’s day. The periods labeled ‘Before CNY’, ‘Extended COVID-lockdown’, and ‘Total COVID-lockdown’ are classified within the first three months of 2020.

During the COVID-19 lockdown period, the NCP region experienced an abnormal haze pollution. To better understand and quantify the contribution of meteorological factors to the haze event, the interval from January 21 to February 9, 2020 was defined as the ‘During-haze’ period; the time before this interval is categorized as ‘Before-haze’ and the time after identified as ‘Post-haze’. Detailed definitions of these period are provided in Table 1.

Table 1

Classification and characteristics of haze-related period in early 2020.

Label	Event Name	Time Span	Characteristics
P1	Before-haze	January 1-January 20, 2020	High emissions; high PM <sub>2.5</sub> level; moderate meteorology
P2	During-haze	January 21-February 09, 2020	Low emissions; high PM <sub>2.5</sub> levels; high RH; moderate T
P3	Post-haze	February 10-February 29, 2020	Low emissions; low PM <sub>2.5</sub> level; Precip; high T; high WS

## 3. Results

### 3.1. Quantify contributions of different drivers for six air pollutants

Fig. 2 demonstrates not only the observed but also the predicted mass concentrations of all six air pollutants during the COVID-19 lockdown. Except for O<sub>3</sub>, the time series of the other air pollutant concentrations all exhibited a fluctuating downward trend in early 2020. This decline is attributed to the related epidemic closure and/or control measures and associated anthropogenic emission reduction of the primary pollutants and the precursors of secondary pollutants (Wang et al., 2021a). In contrast, the continued increase in O<sub>3</sub> could be attributed to the reduction of NO<sub>x</sub> in the VOC-limited O<sub>3</sub> chemistry regime, which results in the O<sub>3</sub> enhancement due to decreased titration by NO (Shen et al., 2022; Wang et al., 2020b). Among six air pollutants, their predictions and the corresponding observations agree with each other well, with the highest PCC of 0.86 for SO<sub>2</sub> and CO, followed by PCC of 0.84 for O<sub>3</sub>, 0.78 for PM<sub>10</sub>, 0.77 for PM<sub>2.5</sub> and 0.76 for NO<sub>2</sub>, the average PCC of which is above the previous studies in China (PCC = 0.71) (Shen et al., 2022) and Spain (PCC = 0.72) (Petetin et al., 2020), indicating robust predictive performance by the ML model. Note that before selecting the optimal ML model, the time series cross validation was applied in four experiments, with the training datasets covering 2015, 2015–2016, 2015–2017,



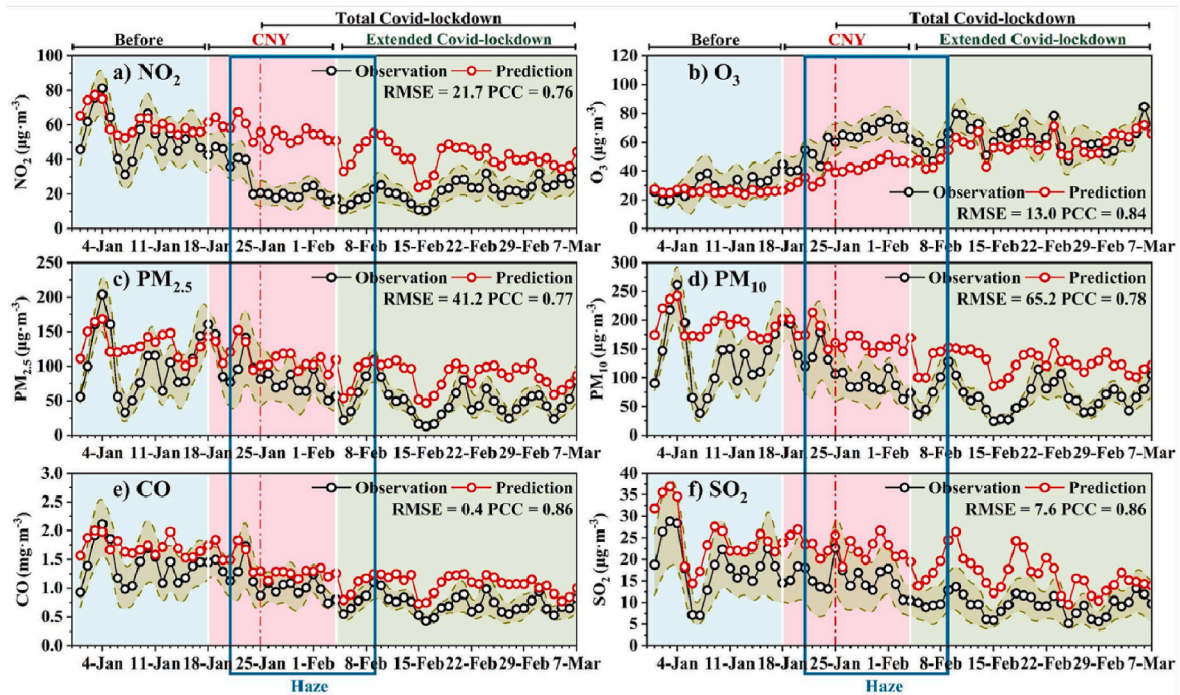


Fig. 2. Time series distribution of observed and predicted daily mean mass concentrations of six air pollutants in Shandong Province. The black circle-line indicates observed concentrations, the red circle-line indicates predicted concentrations and the blue rectangle indicates the haze event in 2020.

2015–2018 and 2015–2019, and the test datasets covering the first three months of 2016, 2017, 2018, 2019 and 2020, respectively. As shown in Fig. S2–S7, six pollutants' mass concentration predictions generally agree with the observations well except for some extreme high or low values generally due to anthropogenic emission changes suddenly.

To evaluate the influence of various anthropogenic drivers, fractional changes in pollutant concentrations between the BAU and factual scenarios were calculated for different periods (Table 2). Before the CNY period, the ML model systematically overestimated the concentrations of pollutants (excluding O<sub>3</sub>) relative to the observed six air pollutants' mass concentrations (Fig. 2). These overestimations likely reflect the effect of local or national air quality improvement regulations, such as the national Clean Air.

Plan in 2018–2020 (CAP<sub>2018–2020</sub>) (Shen et al., 2022). On average, these measures have led to a reduction of −31.1 % for PM<sub>10</sub>, followed by reductions of −26.4 %, −23.5 %, −18.8 % and −11.7 % for SO<sub>2</sub>, PM<sub>2.5</sub>, CO and NO<sub>2</sub>, respectively. For O<sub>3</sub>, the reduction of O<sub>3</sub> chemistry precursors, especially NO<sub>x</sub>, attributable to CAP measures could trigger increased O<sub>3</sub> production (12.6 %) due to its nonlinear response characteristic of VOC-limited photochemical regimes (Zhang et al., 2022). At the beginning of the CNY period, a larger gap between observed and predicted concentrations of NO<sub>2</sub>, SO<sub>2</sub>, PM<sub>10</sub> and O<sub>3</sub> started to appear due to the CNY effect. The above gaps further to widen until the end of the CNY period attributable to concurrent CNY effect and COVID-19 lockdown effect. However, this widening gap was not observed for PM<sub>2.5</sub> in

the same period, indicating the presence of additional production of PM<sub>2.5</sub> driven by specific meteorological conditions (Fig. 2c). During the whole CNY period, the reduction portion of NO<sub>2</sub> (−51.3 %) was higher than that (−47.4 %) in the extended COVID-lockdown period due to the compound effect of the CNY related and the specific COVID-19 lockdown related reductions. In response to the biggest NO<sub>2</sub> reduction during this special period, O<sub>3</sub> production became more intensive and led to a much higher increase (51.2 %) in O<sub>3</sub> concentration than that (10.0 %) in the extended COVID-lockdown period (Sun et al., 2020). By contrast, reductions in PM<sub>2.5</sub> (−21.7 %), PM<sub>10</sub> (−31.8 %), SO<sub>2</sub> (−32.7 %) and CO (−19.7 %) during the CNY period were relative modest and less than those during the extended COVID-lockdown period (PM<sub>2.5</sub>: 43.8 %; PM<sub>10</sub>: 46.3 %; CO: 32.2 %; SO<sub>2</sub>: 42.1 %). Different magnitudes of reduction in PM, CO and SO<sub>2</sub> in the above three periods could not only be attributed to the primary emission reductions but also to secondary formation process such as the gas-to-particle conversions (Chen et al., 2020). These processes are known to be enhanced under stagnant meteorological conditions (Sun et al., 2020).

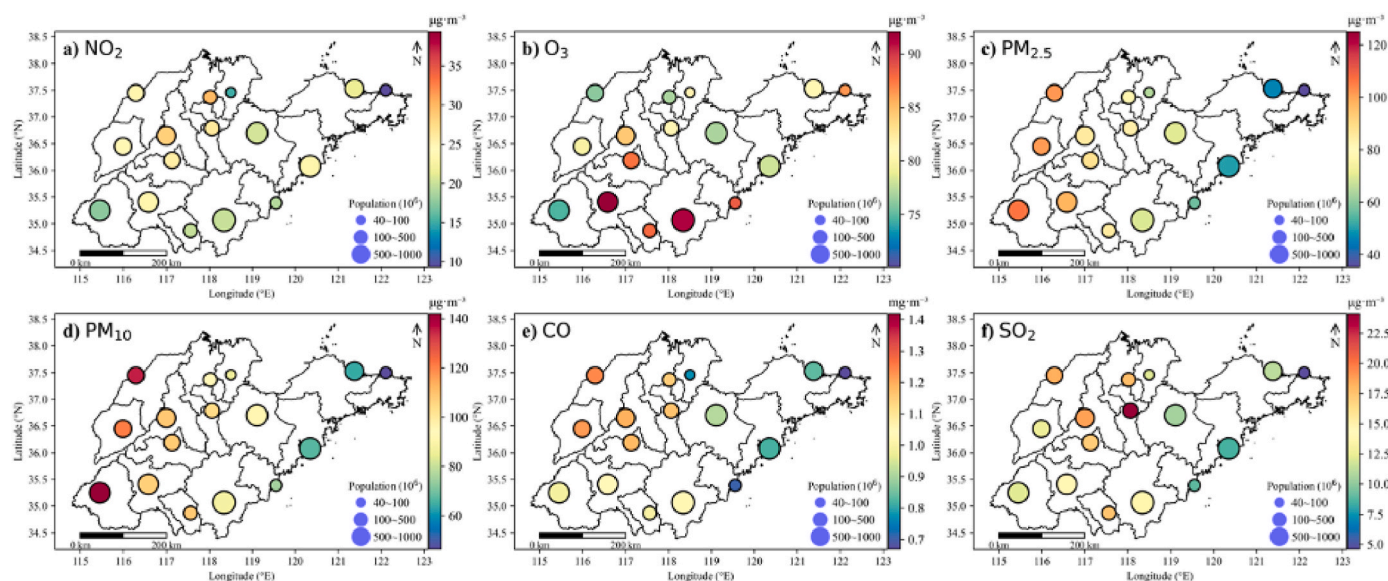
To better investigate the air quality situation during the haze event, the spatial distribution of each pollutant was investigated in Shandong Province. As shown in Fig. 3, except for O<sub>3</sub>, the mass concentrations of air pollutants were higher in the western regions compared to the coastal areas. The average PM<sub>2.5</sub> concentration across Shandong was 76.2 µg/m<sup>3</sup>, with particularly elevated levels observed in Dezhou (101.8 µg/m<sup>3</sup>), Liaocheng (101.5 µg/m<sup>3</sup>), and Heze (106.3 µg/m<sup>3</sup>), where values exceeded 100 µg/m<sup>3</sup>. Similarly, the average PM<sub>10</sub> concentration in Shandong was 96.7 µg/m<sup>3</sup>, with the highest record in Heze at 142.0 µg/m<sup>3</sup>. In contrast, elevated O<sub>3</sub> concentrations were found mainly concentrated in southern Shandong, with the highest value (92.1 µg/m<sup>3</sup>) in Jining. Notably, during the period of the haze event in Shandong, only the average mass concentration of PM<sub>2.5</sub> exceeded the China Ambient Air Quality Standards Grade II (CAAQS-II) threshold of 75 µg/m<sup>3</sup> on average, while the levels of other pollutants remained below their corresponding CAAQS-II thresholds. These results underscore the severity of PM<sub>2.5</sub> pollution in Shandong Province during this period, particularly in the inland areas.

Table 2  
The COVID-lockdown effects in different time periods.

Pollutants	Before CNY	CNY	Extended COVID-lockdown	Total COVID-lockdown
NO <sub>2</sub>	−11.7 %	−51.3 %	−47.4 %	−52.1 %
O <sub>3</sub>	12.6 %	51.2 %	10.0 %	18.9 %
PM <sub>2.5</sub>	−23.5 %	−21.7 %	−43.8 %	−40.0 %
PM <sub>10</sub>	−31.1 %	−31.8 %	−46.3 %	−45.5 %
CO	−18.8 %	−19.7 %	−32.2 %	−29.4 %
SO <sub>2</sub>	−26.4 %	−32.7 %	−42.1 %	−38.7 %

$$\text{Effect (\%)} = ((\text{Obs} - \text{Pre})/\text{Pre}) \times 100.$$



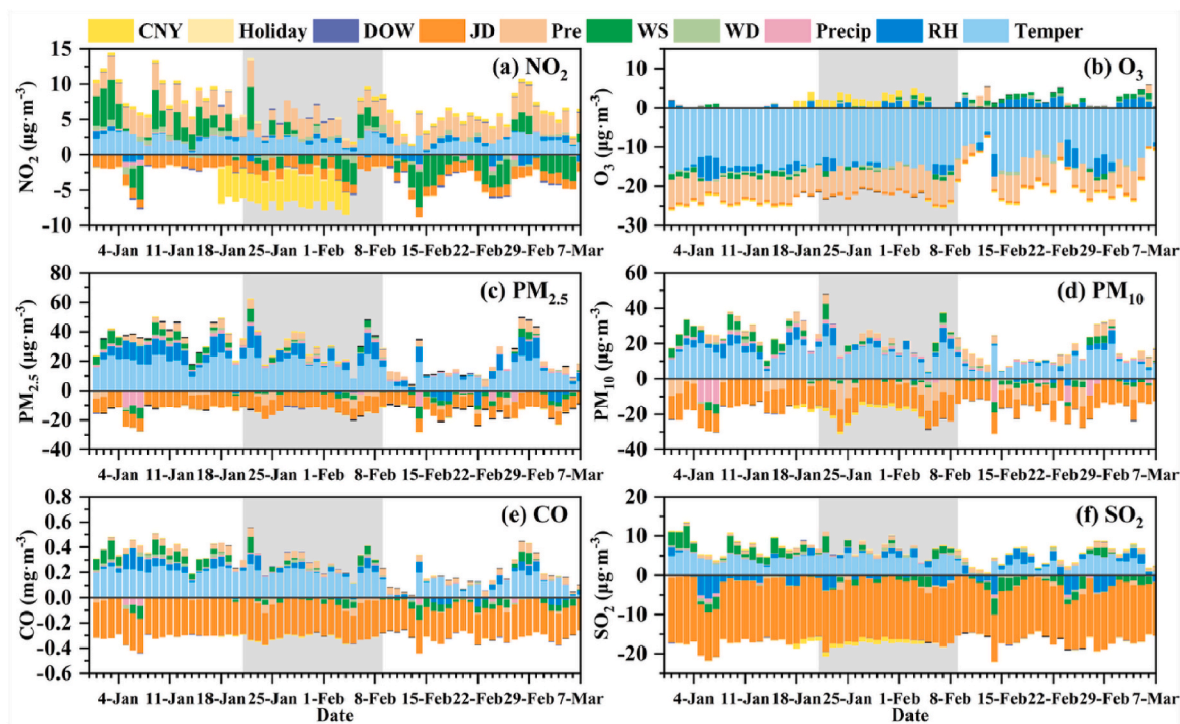


**Fig. 3.** The spatial distributions of observed mean values of six air pollutants during the haze event in 2020. The size of blue solid circle indicates the population size in Shandong Province.

### 3.2. Quantify contributions of meteorological factors for air pollutants

Following the evaluation of the contributions of various drivers for six air pollutants, a SHAP analysis was employed to quantify the importance of the model's input features in the BAU scenario. Fig. 4 shows the time series of SHAP values for all ML model's input features during the COVID-19 lockdown. Notably, time variables (JD, DOW, Holiday, CNY) were used as independent predictors. The time predictors were primarily used as proxies for emissions: the year reflects changes in emission factors, while JD, DOW, Holiday and CNY capture emission trends a day, a week, a holiday, and CNY days respectively (Wang et al.,

2020b). Overall, the CNY variable showed a negative mean SHAP value for  $\text{NO}_2$  ( $-3.0 \mu\text{g}/\text{m}^3$ ), indicating the driver of the CNY holiday plays a dominant role in  $\text{NO}_2$  reduction during the haze period (Fig. 4a). During the Total COVID-lockdown, in terms of negative SHAP, JD was the most significant contributor to  $\text{PM}_{2.5}$  ( $-9.7 \mu\text{g}/\text{m}^3$ ),  $\text{PM}_{10}$  ( $-12.9 \mu\text{g}/\text{m}^3$ ), CO ( $-0.3 \mu\text{g}/\text{m}^3$ ) and  $\text{SO}_2$  ( $-14.5 \mu\text{g}/\text{m}^3$ ) among all the features, reflecting the ongoing daily emission reductions of those four pollutants in Shandong Province (Fig. 4c–f). For  $\text{O}_3$ , the SHAP value of temperature ( $-13.6 \mu\text{g}/\text{m}^3$ ) was the highest, followed by pressure ( $-5.1 \mu\text{g}/\text{m}^3$ ), RH ( $-0.4 \mu\text{g}/\text{m}^3$ ), WS ( $-0.1 \mu\text{g}/\text{m}^3$ ) and WD ( $-0.1 \mu\text{g}/\text{m}^3$ ) in Shandong (Fig. 4b). It reveals that those meteorological factors push  $\text{O}_3$  prediction



**Fig. 4.** Time series distribution of daily SHAP values of input variables (colored bar) for six air pollutants during the COVID-19 lockdown in Shandong, respectively (Note that the shaded area indicates the haze event period).

to a lower value, suggesting that these variables suppressed  $O_3$  toward the baseline levels in Shandong (Fig. 2b). In contrast, SHAP values of temperature for  $PM_{2.5}$  ( $16.5 \mu\text{g}/\text{m}^3$ ),  $PM_{10}$  ( $13.2 \mu\text{g}/\text{m}^3$ ),  $SO_2$  ( $3.9 \mu\text{g}/\text{m}^3$ ) and CO ( $0.2 \mu\text{g}/\text{m}^3$ ) were still ranked the first among the six meteorological factors but turned to positive, indicating its role in promoting pollutant accumulation (Fig. 2c–f).

To better understand meteorology's impact rather than emission reductions on haze formation, we further compared the mean SHAP values of each meteorological factor for  $PM_{2.5}$  between during haze (P2) and after haze periods (P3) (Fig. 5). It should be noted that P2 and P3 coincided with first-level emergency response in Shandong Province. On average, Shandong Province experienced an obvious  $PM_{2.5}$  pollution phase with a value of  $76.3 \mu\text{g}/\text{m}^3$  in P2 and a clean phase with a value of  $45.6 \mu\text{g}/\text{m}^3$  in P3, respectively (Fig. 6). During P2, the temperature had a positive SHAP value for  $PM_{2.5}$  ( $+19.7 \mu\text{g}/\text{m}^3$ ), followed by RH ( $+5.8 \mu\text{g}/\text{m}^3$ ), Precip ( $+0.9 \mu\text{g}/\text{m}^3$ ), WD ( $+0.3 \mu\text{g}/\text{m}^3$ ), Pres ( $+0.1 \mu\text{g}/\text{m}^3$ ) and WS ( $+0.1 \mu\text{g}/\text{m}^3$ ), respectively (Fig. 5c). In comparison, the SHAP values for meteorological factors in P3 showed a significant difference, with lower values for temperature ( $+11.3 \mu\text{g}/\text{m}^3$ ) and RH ( $+1.6 \mu\text{g}/\text{m}^3$ ), a higher value for Pres ( $+1.8 \mu\text{g}/\text{m}^3$ ), and opposite values for Precip ( $-0.6 \mu\text{g}/\text{m}^3$ ), WD ( $-0.1 \mu\text{g}/\text{m}^3$ ) and WS ( $-1.7 \mu\text{g}/\text{m}^3$ ), respectively (Fig. 5c). For temperature, the mean value in P2 ( $1.7^\circ\text{C}$ ) is lower than that in P3

( $5.1^\circ\text{C}$ ), contributing to a lower planetary boundary layer height (PBLH) in P2, which facilitates the accumulation of  $PM_{2.5}$  (Fig. 6f). For precipitation and wind, higher precipitation (P3:  $1.6 \text{ mm}$  vs P2:  $0.1 \text{ mm}$ ) and stronger winds (P3:  $2.8 \text{ m/s}$  vs P2:  $2.4 \text{ m/s}$ ) in P3 promoted wet deposition and transportation, thereby leading to lower concentration of  $PM_{2.5}$  (Fig. 6b–d).

### 3.3. Analyses disturbance of meteorological conditions during haze event

Weather systems played a critical factor in driving the haze event in Shandong Province during the COVID-19 lockdown. To elucidate their influence, we further compare the weather systems between the P2 and P3 periods, focusing on how temporal and spatial variations in atmospheric circulation and physical fields affect haze formation and dissipation (Figs. 7 and 8). At 850 hPa level (Fig. 7a–b), RH exhibited significant difference in southwestern Shandong, where RH decreased from 60 to 70 % in P2 to 50–60 % in P3. Similarly, across the 925–1000 hPa layers (Fig. 7c–f), RH declined by approximately 10 % in P3 relative to P2, indicating a province-wide reduction in atmospheric moisture. This decrease in RH highlights a pronounced decline in humidity conditions from P2 to P3, which likely limited secondary aerosol formation, contributing to the observed decrease in  $PM_{2.5}$ . During P2, the MLSP

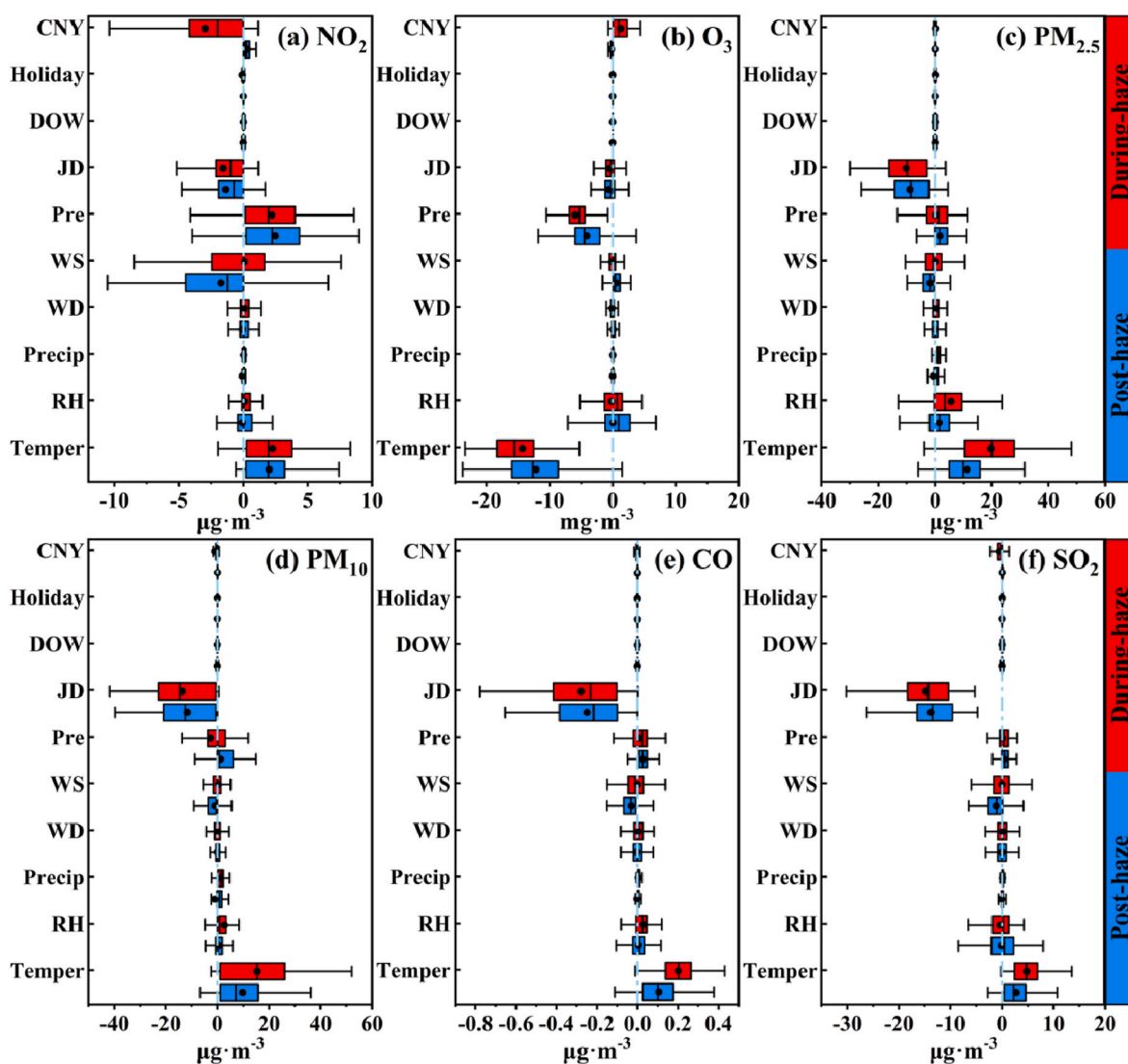


Fig. 5. The box-and-whisker plots represent the mean SHAP value of the input variables for six air pollutants during the haze (red color) and post haze period (blue color) in Shandong Province.



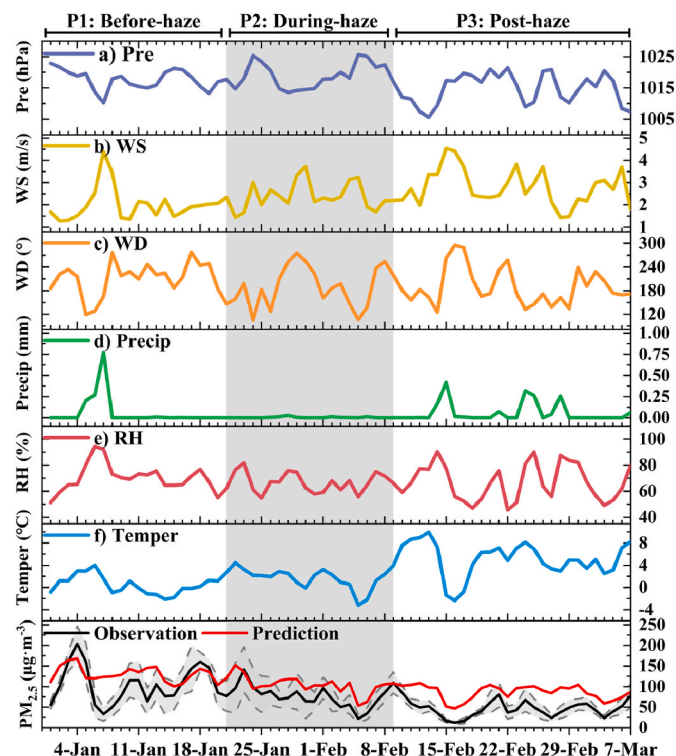


Fig. 6. Time series distribution of observed mean values of six meteorological factors and  $PM_{2.5}$  during the COVID-19 lockdown period (Note that the shaded area indicates the haze event period).

field revealed a dominant high-pressure system over the NCP region, northeastern China, and Shandong Province. Shandong Province, located at the base of this high-pressure system, experienced strong horizontal divergence, which suppressed vertical atmospheric motion and trapped pollutants near the surface (Fig. 7g). In contrast, P3 saw a transition to a broader uniform-pressure environment (Fig. 7h), weakening atmospheric stability and providing more favorable conditions for pollutant dispersion. Notably, the temperature-dew point difference ( $T - T_d$ ) approached zero in P3 (Fig. 7h), primarily due to precipitation-induced saturation, which facilitated the removal of  $PM_{2.5}$  from the

atmosphere. For surface wind field, inland Shandong in P2 experienced persistently low wind speeds, with near-calm conditions dominating in the southwestern region (Fig. 7g). These surface-stagnant meteorological conditions in P2, relative to P3, significantly hindered horizontal pollutant dispersion, leading to the accumulation of particulate matter. In terms of temperature from P2 to P3, as shown in Fig. 8, it increased by  $2^\circ\text{C}$ – $4^\circ\text{C}$  at the 850 hPa level and at the surface (2 m above ground), enhancing PBLH and improving vertical dispersion conditions, which contributed to better surface air quality. In addition to the influence of local meteorological factors on  $PM_{2.5}$  pollution in P2, regional pollution transport also played a crucial role. This external pollution influx in P2 was driven by a pronounced warm ridge over the NCP region (Fig. 8a) coupled with prevailing northwesterly winds at 850–925 hPa (Fig. 7a–c), resulting in the transport of pollutants from the NCP region to Shandong.

Overall, specific weather patterns under the impact of different weather elements concurrently were attributed to the  $PM_{2.5}$  accumulation and dispersion in Shandong Province during P2 and P3 periods. These results consistently highlight the critical role of weather systems in haze formation and provide helpful insights into meteorological modulation in air pollution in Shandong Province.

## 4. Discussions

### 4.1. Anthropogenic drivers for the air pollution change during the COVID-19 lockdown period

Regarding the anthropogenic contributors, the COVID-19 lockdown measures, the CAP<sub>2018-2020</sub> policy, and the CNY holiday all made a certain contribution to the change in the major pollutants. Previous studies have demonstrated that both routine and special societal slowdowns, such as the CNY holiday, CAP implementation, and high-profile events like the Olympic Games and G20 Summits, are typically associated with decreased industrial and vehicular activity, subsequently leading to reductions in pollutant emissions (Pei et al., 2020). For instance, Wang et al. (2010) estimated that over 60 % of  $PM_{2.5}$  reduction during the 2008 Olympics could be attributed to emissions control strategies. Similarly, Yan et al. (2012) reported lower near-surface  $SO_2$  concentrations in the Pearl River Delta during the 2010 Asian Games. Huang et al. (2015) found a 47 % reduction in tropospheric  $NO_2$  concentrations over the North China Plain during the 2014 APEC meeting

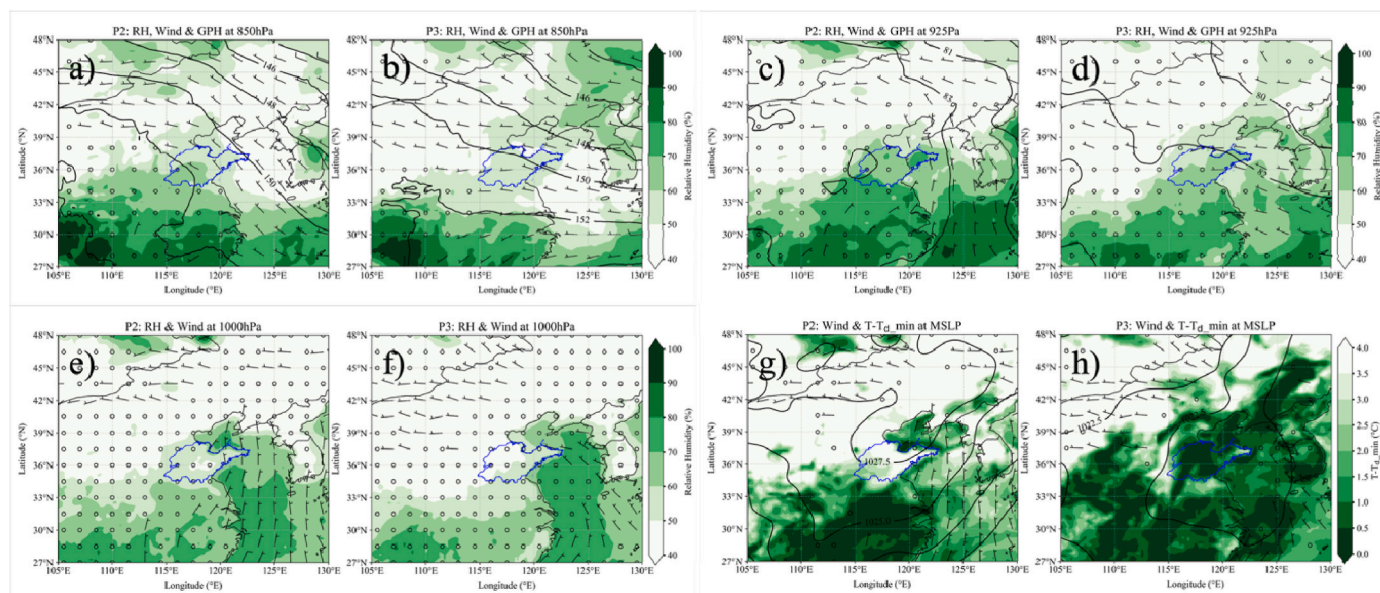


Fig. 7. Comparison of the RH and wind fields at 850 hPa, 925 hPa, 1000 hPa levels and  $T-T_d$  field at MSLP between During-haze and Post-haze periods, respectively.



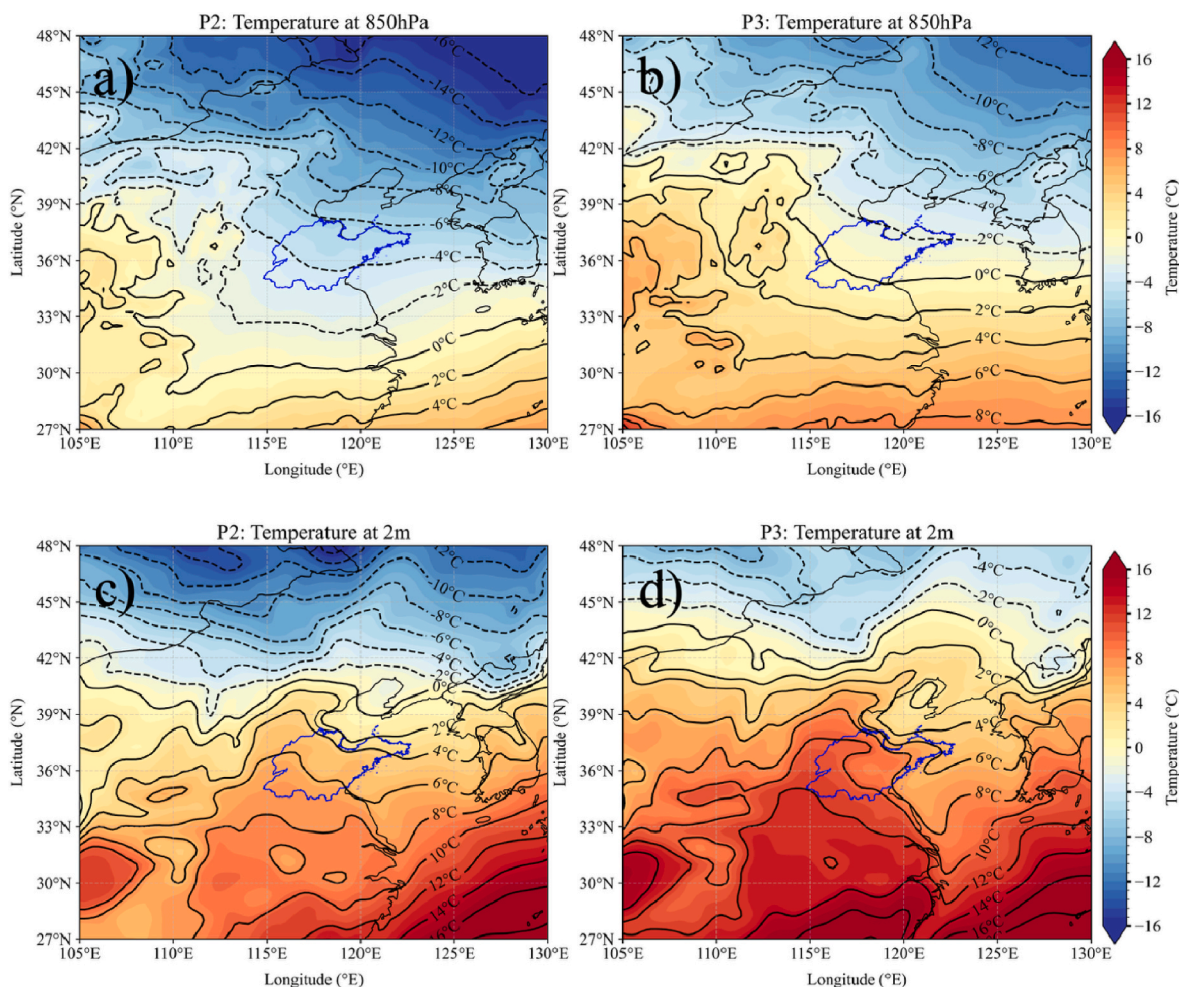


Fig. 8. Comparison of temperature fields at 850 hPa and 1000 hPa levels between During-haze and Post-haze periods, respectively.

relative to the same period in the preceding three years. Li et al. (2016) showed a 52–57 % reduction in  $PM_{10}$  mass concentration during the 2015 Victory Day Parade after controlling for meteorological variability.

Over the past decade, a major routine anthropogenic intervention has been China's Clean Air policy initiatives. In 2013, the Chinese State Council launched a five-year Air Pollution Prevention and Control Action Plan (referred to as the "Ten Measures of Air"), targeting  $PM_{2.5}$  reduction by the end of 2017 (Shen et al., 2020). Building on this, the government issued the "Three-Year Action Plan to Win the Blue Sky Defense War" in 2018 to address  $PM_{2.5}$  and  $O_3$  pollution jointly (Shen et al., 2022). Following implementation of these policies, concentrations of  $PM_{2.5}$ ,  $PM_{10}$ ,  $NO_2$ ,  $SO_2$ , and CO decreased by 46.2 %, 43.1 %, 24.7 %, 55.5 %, and 37.2 %, respectively. In contrast, 8-h  $O_3$  concentrations increased by 19.4 %, highlighting the complexity of secondary pollutant formation (Li et al., 2023).

The CNY holiday represents another routine social pause with moderate emission impacts. This social pause generally starts one week before the Chinese New Year Day and lasts two weeks or more (Miyazaki et al., 2020). Although industrial activity declines during this period, many manufacturing operations continue, and traffic flows shift—often moving from urban centers to rural areas (Guo et al., 2018). Unlike these routine events, the COVID-19 lockdown was an unprecedented, nationwide cessation of both industrial production and transportation for over a month (Tian et al., 2020). This unique event offered a rare opportunity to evaluate baseline atmospheric conditions in the absence of typical anthropogenic influences. Numerous studies have since reported heterogeneous air quality responses across regions during the

lockdown (Nie et al., 2021; Shen et al., 2024; Wang et al., 2021a). Our findings presented here underscore the value of continued investigation into the differential impacts of anthropogenic activity on atmospheric composition.

#### 4.2. Influence of meteorological input features on haze event

To better understand the influence of meteorological conditions on haze formation by SHAP in Shandong Province, the contribution of each input feature to local  $PM_{2.5}$  concentration predictions was examined while accounting for intra-model variability. Here, intra-model variability of SHAP values refers to differences in SHAP attributions for the same feature value across different phases of a haze event, reflecting the complex mechanisms governing  $PM_{2.5}$  formation and removal.

During phase P2, the highest average SHAP value for ambient temperature indicated a strong positive association with  $PM_{2.5}$  concentrations, followed by RH (Fig. 5c). The statistical model successfully reproduced established patterns linking  $PM_{2.5}$  species variability to temperature. For instance, sulfate formation is known to increase at higher temperatures due to enhanced oxidation rates of  $SO_2$  (Li et al., 2017). Chen et al. (2025), using SHAP analysis, suggested that temperature is the dominated meteorological factors influencing  $PM_{2.5}$  levels through changing the atmospheric oxidation capacity in Beijing-Tianjin-Hebei (BTH), China. In this study, the higher  $PM_{2.5}$  levels observed in P2 at lower average temperatures (1.7 °C) compared to P3 (5.1 °C) are inconsistent with this phenomenon. This pattern is likely related to enhanced aqueous-phase sulfate formation under colder

and foggier conditions in Beijing (Wang et al., 2020a). Similarly, Stirnberg et al. (2021) reported other potential mechanisms at lower temperatures aligned with the SHAP results, such as the significant nitrate formation and the enhanced organic matter, which are the two well-documented processes (Petit et al., 2015; Wang et al., 2021b).

The relatively higher RH SHAP value for PM<sub>2.5</sub> also played a positive role in the high PM<sub>2.5</sub> levels during P2. RH could influence the formation mechanisms of nitrate and sulfate by promoting heterogeneous reactions and serves as an indicator of those occurrence (Ge et al., 2012; Sun et al., 2013). Liu et al. (2025), in a study of exploring the drivers of PM<sub>2.5</sub> pollution by the SHAP method, have suggested that SO<sub>2</sub> oxidation could be accelerated in a condition of RH exceeds 50 % but suppressed with RH above 90 %, likely due to the rainfall-induced scavenging of PM<sub>2.5</sub>. The observed difference in SHAP value for PM<sub>2.5</sub> between P2 and P3 is consistent with this mechanism. In P2, the lower average RH (66.6 %) with respect to P3 likely lead to more PM<sub>2.5</sub> formation, as several rainfall events occurred in P3 (Fig. 6d).

#### 4.3. Limitations

This study quantified the contributions of anthropogenic drivers and meteorological factors to pollutant reductions using machine learning and SHAP analysis. While, there are several limitations in disentangling anthropogenic drivers.

First, the treatment of anthropogenic drivers in this study was indirect and primarily represented by time-related variables (e.g., holidays, Chinese New Year, Julian day). Although these proxies can capture variations in human activity, they do not provide direct information about changes in specific emission sources or anthropogenic behaviors. As a result, the quantification of anthropogenic drivers may be less precise than the assessment of meteorological factors. Moreover, the study did not explicitly categorize anthropogenic drivers by emission sector (e.g., industry, transportation, residential combustion). The reliance on a residual-based approach derived from machine learning predictions makes the interpretation of anthropogenic contributions more abstract. Finally, the lack of direct emission data or more detailed quantitative indicators of social and economic activities limits the ability to fully separate and attribute anthropogenic effects. Future studies could address this limitation by designing a framework that integrates multi-data fusion analysis (e.g., detailed emission inventories, sectoral activity data, or real-time indicators of anthropogenic activity) using an ensemble machine learning method to improve attribution accuracy.

#### 5. Conclusion

This study evaluates the contributions of different drivers and meteorological factors to the changes in six air pollutants at 16 surface sampling sites in Shandong Province during the COVID-19 lockdown. From ML results, it was found that the combined effect of the Total COVID-lockdown measures and regular CNY celebration contributed to an overall 52.1 % reduction in NO<sub>2</sub>, followed by reductions of 45.5 %, 40.0 %, 38.7 %, and 29.4 % for PM<sub>10</sub>, PM<sub>2.5</sub>, SO<sub>2</sub> and CO, respectively. In contrast, an increase of 18.9 % in O<sub>3</sub> was observed during the Total COVID-lockdown in Shandong. Based on SHAP values and the analysis of meteorological disturbances, we found that temperature, with the SHAP values of 19.7 µg/m<sup>3</sup> in P2, was the dominant meteorological factor in the weather system, facilitating the formation of the haze event during the COVID-19 lockdown period in Shandong.

Overall, this study provides robust insights into the impact of meteorological factors on air pollution by quantifying their contributions during the COVID-19 lockdown in Shandong Province. By analyzing the weather patterns during the same period, it further proves the accuracy and effectiveness of the ML model and SHAP module, which aid in the development of more effective air quality management strategies. This approach highlights the potential for using data-driven

methods to design more effective policies for improving air quality and safeguarding public health. However, in order to improve the performance of the ML model, our work needs to be extended to include other related meteorological factors.

#### CRedit authorship contribution statement

**Yue Yuan:** Writing – review & editing, Writing – original draft, Visualization, Validation, Software, Resources, Project administration, Funding acquisition, Formal analysis, Data curation, Conceptualization. **Fuzhen Shen:** Writing – review & editing, Supervision, Methodology, Formal analysis, Data curation, Conceptualization. **Chunyan Sheng:** Software, Resources, Investigation, Formal analysis. **Zeming Zhang:** Software, Resources, Formal analysis, Data curation. **Weihua Guo:** Visualization, Software, Resources, Data curation. **Wengang Zhu:** Visualization, Software, Resources, Formal analysis. **Hui Zhu:** Visualization, Software, Resources, Formal analysis.

#### Code availability

The python code and data used to run the LightGBM model can be obtained upon request from the corresponding authors.

#### Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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#### Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.apr.2025.102694>.

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