

Integrating Deep Learning and Multi-Objective Optimization for Floodwater Utilization: A Coordinated Surface Water-Groundwater Regulation Framework for Groundwater Recovery

Libin Zhang^a, Yonggen Zhang^{b*}, Jianzhu Li^a, Shaowei Lian^c, Xinwang Li^d, Ping Feng^a, Lutz Weihermüller^e

^a State Key Laboratory of Hydraulic Engineering Intelligent Construction and Operation, Tianjin University, Tianjin 300072, China

^b Institute of Surface-Earth System Science, School of Earth System Science, Tianjin University, Tianjin 300072, China

^c Hebei Provincial Hydrologic Survey and Research Center, Shijiazhuang 050051, China

^d Hebei Institute of Water Science (Hebei province dam safety technology center, Hebei province levee sluice technology center), Shijiazhuang 050051, China

^e Agrosphere Institute IBG-3, Forschungszentrum Jülich GmbH, Jülich, 52428, Germany

*Correspondence to: Yonggen Zhang (ygzhang@tiu.edu.cn)

17 **Abstract:** Climate change has exacerbated global droughts and floods, further
18 disrupted the uneven temporal and spatial distribution of water resources, and therefore,
19 poses a significant challenge to water resource management. Flood utilization,
20 converting floodwater from hazard to valuable resource, is a key solution to this
21 challenge. However, existing flood utilization strategies predominantly focus on
22 surface water management through reservoir operations, overlooking integrated
23 optimization with groundwater systems, particularly the challenges of coupling
24 physical models with multi-objective algorithms for groundwater recovery. Here, by
25 ensuring ecological flow and downstream flood safety, a multi-objective optimization
26 framework employing deep learning was developed to integrate flood control, water
27 storage, and groundwater recovery. Reservoir operations were optimized through multi-
28 scenario simulations, and a 3D groundwater numerical model was employed to assess
29 the impact of managed aquifer recharge (MAR) using floodwater on groundwater
30 recovery. Results for the 2023 flood season (June to September) showed that,
31 increasing the flood limited water level (FLWL) reduced average reservoir flood risk
32 and water scarcity by 84.9% and 61.9%, respectively, while weakening their inverse
33 relationship. This indicates that raising FLWL improves individual objectives and
34 reduces conflicts for balanced optimization. Maintaining continuous ecological river
35 flow promoted groundwater recovery despite reduced total river discharge. MAR at 300
36 $\text{m}^3 \text{ d}^{-1}$ achieved effective groundwater recovery in 17.6% of the study area with a
37 maximum of 0.46 m. Overall, this study presents a novel framework coupling deep
38 learning, multi-objective optimization, and 3D groundwater modeling, enabling

39 optimized surface water-groundwater regulation and enhanced floodwater utilization
40 for groundwater recovery.

41 **Keywords:** Floodwater utilization; Deep learning; Groundwater recovery; Multi-
42 objective optimization; Ecological flow; Managed aquifer recharge

43 **1. Introduction**

44 Global climate change is expected to intensify the hydrological cycle ([Tabari, 2020](#)),
45 resulting in an increase in extreme precipitation events ([Blöschl et al., 2017](#)), leading
46 to larger flood frequencies ([Chagas et al., 2022](#)), and posing significant challenges to
47 water resource management, particularly in monsoon regions ([Hirabayashi et al., 2013](#);
48 [Yang et al., 2023](#)). This increase in flood intensity and frequency poses threats to
49 ecosystems, economies, and human livelihoods ([Bermúdez et al., 2021](#)). In response to
50 these challenges, innovative approaches such as floodwater utilization (FU) have been
51 developed, which converts floodwater from a hazard into a valuable resource ([Wang et](#)
52 [al., 2023](#)). This strategy has been increasingly adopted in integrated river basin
53 management to balance water allocation ([Li et al., 2021](#)) and reservoir operation
54 optimization to mitigate flood risks and water scarcity ([Jiang et al., 2019](#)), and enhance
55 storage and recharge, while simultaneously controlling flood risks through strategic
56 reservoir operations and flow regulation.

57 FU strategies typically involve drawing down reservoirs before the flood season to
58 create additional storage capacity, capturing and storing excess floodwater during the
59 flood season, and releasing the stored water for beneficial purposes such as irrigation
60 and hydropower generation during the non-flood season ([Wang et al., 2023](#)).
61 Maximizing reservoir operation benefits was achieved through multi-stage ([Liu et al.,](#)
62 [2015; Wei et al., 2022](#)) or dynamic flood limited water level (FLWL) adjustments ([Ding](#)
63 [et al., 2023](#)). Furthermore, the joint operation of multiple reservoirs ([Jain et al., 2023](#))
64 and the strategic use of flood retention areas ([Bellu et al., 2016](#)) have proven effective

65 in optimizing FU. Where applied, these measures have reduced flood damage and
66 enhanced floodwater utilization efficiency ([Mateo et al., 2014](#)).

67 However, while these FU strategies primarily focus on surface water management,
68 research on the integrated optimization of groundwater systems, a fundamental
69 component of the hydrological cycle ([Irvine et al., 2024](#)), remains limited. As critical
70 hydrological components, groundwater resources undergo accelerated global depletion,
71 threatening ecosystems and livelihoods ([Jasechko et al., 2024](#)). This is particularly
72 severe in regions such as the North China Plain, which is one of the world's largest
73 groundwater depression cones ([Chen et al., 2020](#)), highlighting the urgency of
74 groundwater recovery efforts to sustain ecosystems and human activities. Consequently,
75 effective groundwater recovery measures are imperative to counteract depletion and
76 sustain ecosystems and livelihoods.

77 Accurate simulation of groundwater dynamics is essential for optimizing resource
78 allocation in integrated water management ([Haaf et al., 2023](#)). Numerical simulation,
79 the most widely used approach for modeling groundwater dynamics, has significantly
80 contributed to analyzing spatial and temporal groundwater changes and quantifying the
81 effects of various measures on groundwater dynamics ([Condon et al., 2021](#)). Roy et al.
82 ([2024](#)) employed a MODFLOW model to determine the optimal groundwater recharge
83 rate and to minimize groundwater decline. Lyu et al. ([2025](#)) employed an enhanced
84 version of the SWAT-MODFLOW model to evaluate the effects of ecological recharge
85 from reservoirs and reclaimed water releases on groundwater recovery. However,
86 physical models might face limitations when integrated with optimization algorithms

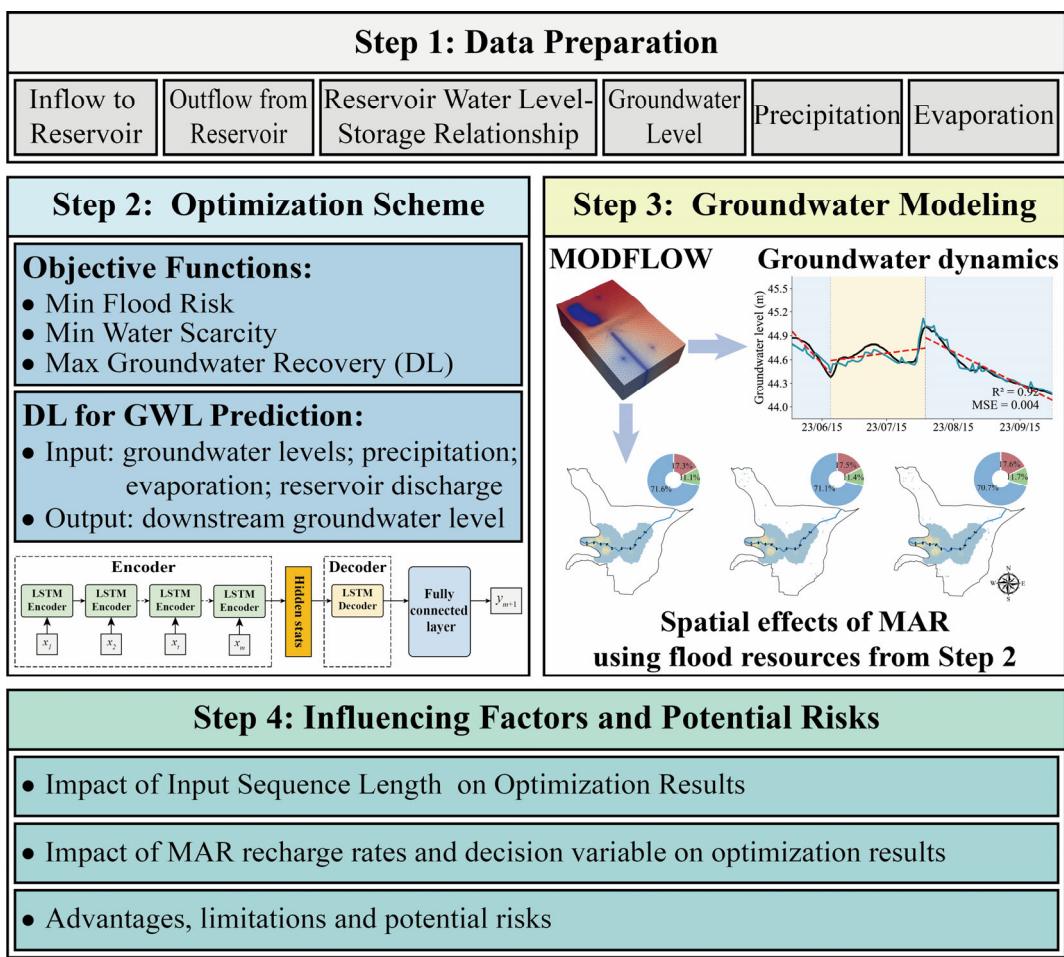
87 owing to their high computational intensity ([Asher et al., 2015](#)), as groundwater
88 dynamics are influenced by multiple hydrogeological factors, including heterogeneity
89 in subsurface hydraulic conductivities and extraction intensities ([Kuang et al., 2024](#)).

90 Recent advancements in deep learning, particularly Long Short-Term Memory
91 (LSTM) networks, have yielded robust computational tools for groundwater prediction
92 ([Tripathy and Mishra, 2024](#)). Due to their predictive accuracy ([Cui et al., 2024](#)) and
93 potential for integration with optimization algorithms ([He et al., 2022](#)), LSTM networks
94 have emerged as a preferred approach for analyzing nonlinear temporal features in
95 hydrological time-series data ([Hochreiter and Schmidhuber, 1997](#)). However, analyzing
96 the impacts of environmental factors on groundwater dynamics necessitates multi-step
97 prediction approaches. To address error accumulation in such multi-step predictions,
98 modifications to the LSTM architecture are necessary to minimize error propagation
99 ([Zhuang et al., 2023](#)). The encoder-decoder LSTM architecture, which can effectively
100 mitigate these issues by capturing temporal dependencies to improve prediction
101 accuracy, has been extensively applied to temporal pattern recognition tasks, including
102 natural language processing and activity forecasting ([Deng et al., 2019](#)). This
103 architecture consists of encoder-decoder modules based on recurrent neural networks.
104 The encoder transforms variable-length sequences into fixed-dimensional context
105 vectors, while the decoder produces predictions by processing these vectors.
106 Comparative analyses demonstrate that the encoder-decoder LSTM framework offers
107 superior forecasting precision compared to conventional deep learning approaches
108 ([Wunsch et al., 2021; Xiang et al., 2020](#)).

109 Despite many studies exploring floodwater utilization (FU) by reservoir regulation
110 and the application of deep learning for groundwater level prediction, few studies have
111 investigated the effects of using flood resources to promote groundwater recovery, by
112 combining groundwater recovery with FU as a framework, leaving a critical gap in
113 addressing the escalating global challenges of climate-driven floods, droughts, and
114 widespread aquifer depletion. To address the limitations of prior studies, this study
115 developed a framework, which embeds a deep learning model to predict groundwater
116 levels, coupled with FU in multi-objective optimization. This framework also used a
117 3D groundwater numerical simulation to evaluate the spatial effects of managed aquifer
118 recharge (MAR) using floodwater resources. The Lincheng Reservoir and its associated
119 downstream groundwater depression cone in the North China Plain served as the
120 testbed for this study.

121 The specific objectives of the study were: (1) to develop an encoder-decoder LSTM
122 model for predicting groundwater levels and integrate it as a component of the objective
123 function in multi-objective optimization; (2) to establish a multi-objective optimization
124 framework that balances flood control, water storage, and groundwater recovery under
125 constraints of ecological flows and flood safety; (3) to assess the effects of managed
126 aquifer recharge (MAR) using floodwater on groundwater recovery through a 3D
127 numerical model; (4) to evaluate the impacts of variations in deep learning
128 hyperparameters (e.g., input sequence length) and MAR recharge rates on the
129 framework's performance and optimization outcomes. The overall technical framework
130 and workflow are illustrated in [Figure 1](#). As depicted in [Figure 1](#), the framework begins

131 with data preparation in Step 1. Building on this, this study integrates deep learning into
 132 multi-objective optimization in Step 2 to jointly optimize flood control, water storage,
 133 and groundwater recovery. Subsequently, in Step 3, a numerical groundwater model
 134 (MODFLOW) is developed to quantify the effects of MAR on groundwater recovery,
 135 thereby transforming floodwater obtained from the previous step into a sustainable
 136 resource for groundwater restoration. In Step 4, this study discusses key factors, such
 137 as input sequence length and MAR rates, that influence the optimization outcomes and
 138 the potential risks involved. This approach provides a methodology for integrated water
 139 resource management in regions with groundwater overexploitation and intensive
 140 agricultural irrigation.



141
 142

Figure 1. Schematic of the technical roadmap developed in this study, integrating

143 an optimization scheme for reservoir flood risk mitigation, water scarcity
144 reduction, and groundwater recovery benefits. The MODFLOW model was used
145 to evaluate the effectiveness of MAR, utilizing flood resources generated by the
146 optimization scheme. The study also investigates key influencing factors and
147 potential risks. Abbreviations in the figure: DL, deep learning; GWL,
148 groundwater level; MAR, managed aquifer recharge.

149 **2. Materials and Methods**

150 **2.1 Study Area**

151 The Lincheng Reservoir is located in Xingtai City, Hebei Province, China and is part
152 of the Ziya River system within the Haihe River basin ([Figure 2a](#)), which lies in a warm
153 temperate continental monsoon climate zone. The average annual temperature ranges
154 from 10 to 13°C, with significant variability in both inter-annual and intra-annual
155 precipitation. Annual precipitation averages 490-600 mm, with 75-80% occurring
156 during the summer months (July-August). Flood season precipitation primarily occurs
157 in the form of high-intensity storms between late July and early August, resulting in an
158 uneven temporal distribution of precipitation concentrated in short, intense periods.

159 The Lincheng Reservoir is primarily designed for flood control and water supply
160 management. With a catchment area of 384 km², its key water levels include the dead
161 storage level (112.0 m), the flood limited water level (FLWL, 120.48 m), and the normal
162 storage level (125.5 m). The discharge from Lincheng Reservoir flows into the Zhi
163 River, which then joins the Fuyang River ([Figure 2b](#)). The Zhi River basin is located
164 within a representative shallow groundwater depression cone (GDC) in the North China
165 Plain, i.e., the Ning-Bai-Long Cone (see the gray shaded area in [Figure 2b](#)), where
166 groundwater levels have undergone significant decline. Since 1980, the development

167 of the GDC has progressed through three primary stages ([Figure 2c](#)). From 1980 to
168 2014, the GDC expanded rapidly, with the average groundwater level experiencing a
169 substantial decline of 30.3 m. Between 2014 and 2018, the expansion of the GDC
170 slowed and the area-wide average groundwater level decreased again by 1.62 m. Since
171 2018, the ongoing implementation of over-extraction control measures has significantly
172 reduced the expansion of the GDC, and as a consequence, the average groundwater
173 level in the GDC had increased by 2.84 m in 2022 compared to 2018.

174 Considering these hydrogeological characteristics and the need for continued
175 management of groundwater depletion, the study focuses on the Lincheng Reservoir
176 and its downstream Zhi River basin, where optimized floodwater utilization offers an
177 effective approach for coordinating multiple water management objectives, including
178 flood control, water storage, ecological flow maintenance, and groundwater recovery.

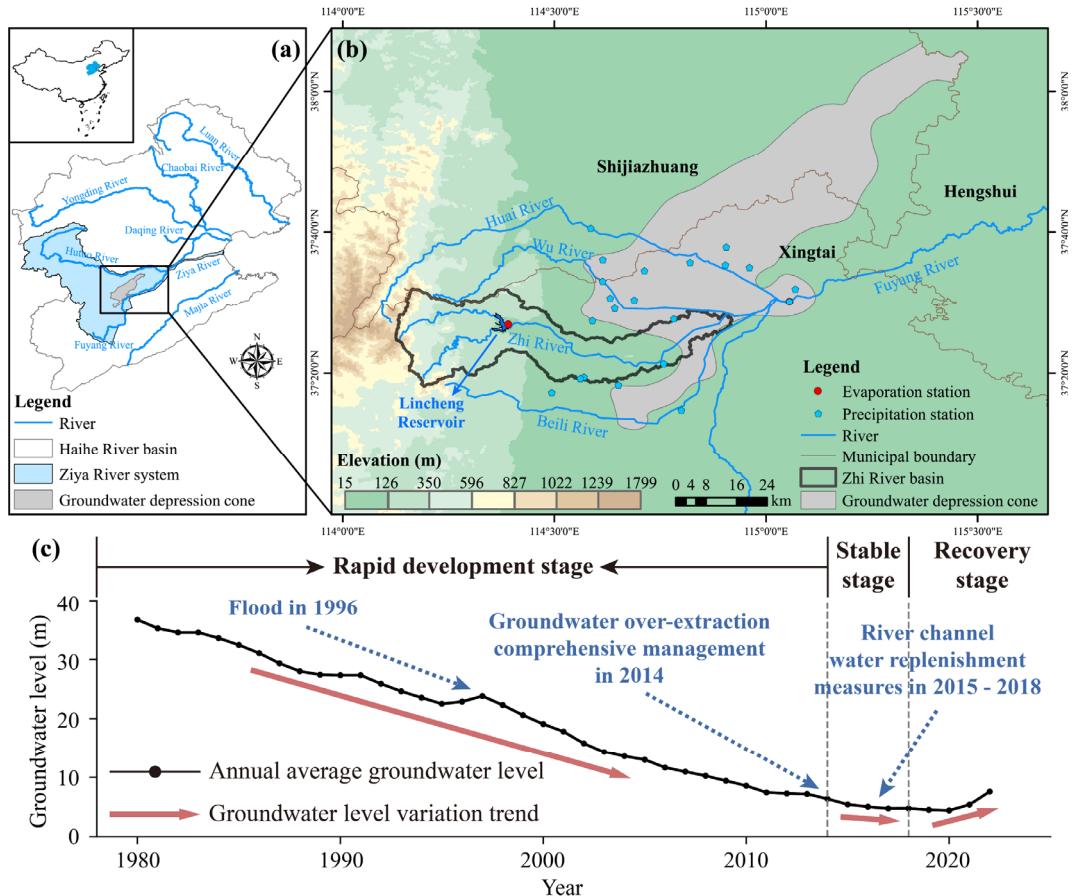


Figure 2. Overview of the study area: (a) Regional setting of the Haihe River basin and Ziya River system, with an inset map illustrating the geographical location of the Haihe River basin in China; (b) Hydrological characterization, including the Lincheng Reservoir, natural rivers, the groundwater depression cone, and hydrological monitoring stations; (c) Temporal evolution of the average groundwater level in the Ning-Bai-Long groundwater depression cone since 1980.

2.2 Data Sources

To support the integrated modeling framework for reservoir operations and groundwater dynamics in the Lincheng Reservoir and Zhi River basin, a comprehensive suite of hydrological, geological, and geospatial datasets was assembled from multiple authoritative sources. Daily inflow and outflow records for the reservoir, critical for simulating water balance and operational scenarios, were provided by the Xingtai Hydrological Survey and Research Center in Hebei Province, China. Meteorological monitoring data, including daily precipitation, potential evaporation, along with

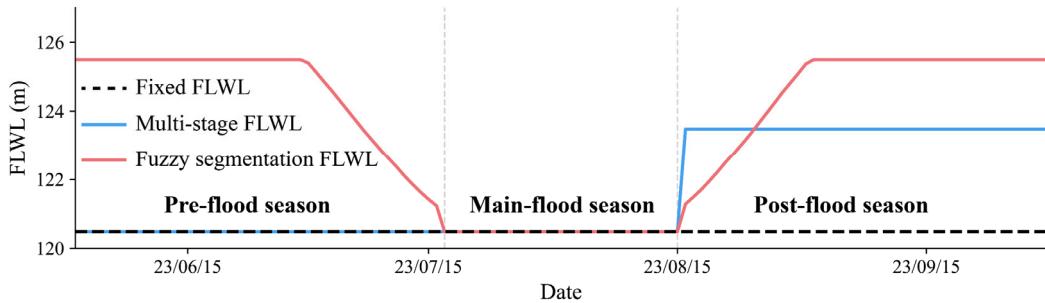
194 groundwater level observations, were sourced from the Hebei Provincial Hydrological
195 Survey and Research Center to capture temporal variability in recharge and depletion
196 patterns. Complementing these, high-resolution geospatial data, including a 90-m
197 Digital Elevation Model (DEM) from the Geospatial Data Cloud
198 (<http://www.gscloud.cn>) and 30-m land-use classifications from the GLC-FCS30
199 dataset (<https://zenodo.org>), enabled accurate delineation of the catchment topography
200 and surface characteristics influencing runoff and infiltration processes. For the 3D
201 groundwater numerical model, key hydrogeological parameters, including geological
202 borehole logs, specific yields, hydraulic conductivities, infiltration coefficients, and
203 groundwater extraction volumes, were obtained from the Ninth Geological Brigade of
204 the Hebei Bureau of Geology and Mineral Resources.

205 **2.3 Methods**

206 **2.3.1 Scenarios Setting of Flood Limited Water Level**

207 To quantify the effects of changes in the flood limited water level (FLWL) on
208 floodwater utilization (FU), this study developed three FLWL scenarios: (a) a fixed
209 FLWL of 120.48 m, (b) a multi-stage FLWL, and (c) a fuzzy-segmentation FLWL,
210 whereby the flood season was divided into pre-flood, main-flood, and post-flood
211 seasons using fuzzy set theory ([Mu et al., 2022](#)). Detailed methods for flood season
212 segmentation and fuzzy segmentation of FLWL determination are provided in the
213 Supporting Information Text S1. Based on this segmentation, the FLWL for scenario
214 (b) was set at 120.48 m for the pre-flood and main flood seasons and 123.48 m for the
215 post-flood season. This configuration was recommended by the Hebei Provincial Water

216 Resources Department, based on their empirical assessments of historical flood patterns
 217 and reservoir safety. Nevertheless, alternative FLWL configurations are feasible within
 218 the proposed framework in this study. In scenario (c) the FLWL for each period was
 219 dynamically adjusted by linking the available flood control storage capacity to the
 220 varying reservoir levels across the pre-flood, main-flood, and post-flood seasons, as
 221 derived from historical precipitation patterns using fuzzy set theory. Detailed
 222 procedures and equations are provided in Supporting Information Text S1. [Figure 3](#)
 223 illustrates the flood season segmentation and FLWL variations across scenarios.



224
 225 **Figure 3. Variations in flood limited water level (FLWL) under different scenarios:**
 226 **the black dashed line represents fixed FLWL (Scenario a), the blue solid line**
 227 **indicates multi-stage FLWL (Scenario b), and the red solid line depicts fuzzy**
 228 **segmentation FLWL (Scenario c).**

229 2.3.2 Groundwater Level Prediction Using encoder-decoder LSTM

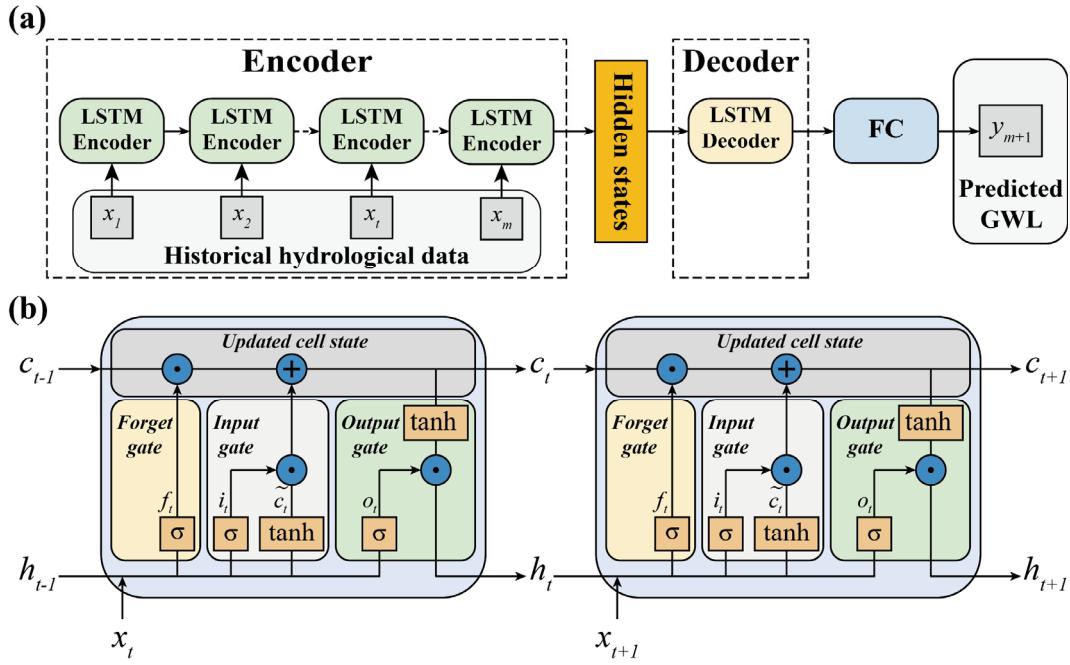
230 This study adopted an encoder-decoder LSTM model, as its architecture serves as an
 231 extension of the LSTM and can better handle longer and more complex input sequences.
 232 In this architecture, the encoder processes a sequence of inputs (such as precipitation
 233 and evaporation) into a summarized representation that captures the essential temporal
 234 information ([Sutskever et al., 2014](#)), and then the decoder uses this information to
 235 predict the target variable (such as groundwater level). Compared to a simple LSTM,
 236 this design enables the model to capture long-term dependencies more effectively and

237 reduces error accumulation during recursive forecasting.

238 The adopted encoder–decoder LSTM framework effectively represents the dynamic
239 interaction between reservoir operations and groundwater responses. As depicted in
240 [Figure 4a](#), the encoder sequentially processes multiple hydrological and meteorological
241 variables over several preceding time steps, capturing their temporal dependencies and
242 compressing the information into a temporary state vector. At each time step t , the input
243 vector x_t comprises historical groundwater levels, precipitation, potential evaporation,
244 and reservoir discharge. After processing an input sequence of length m days, the
245 encoder outputs the final hidden and cell states (h_m, c_m) , which encapsulate the temporal
246 dependencies of the sequence. The decoder then initializes with these states and
247 produces the predicted groundwater level at the next time step (y_{m+1}) through a fully
248 connected layer.

249 Both the encoder and decoder are composed of LSTM units that share the same
250 internal gate structure and information-update mechanism. As shown in [Figure 4b](#), the
251 LSTM unit consists of three gates: the forget gate, input gate, and output gate. The
252 forget gate (f_t) determines the proportion of the previous cell state c_{t-1} to discard,
253 allowing the model to remove outdated information. Subsequently, the input gate (i_t)
254 regulates the current external information (i.e., input x_t) and generates candidate cell
255 state \tilde{c}_t that represents potential new memory. The updated cell state c_t combines the
256 past memory, weighted by the forget gate, with the candidate memory, weighted by the
257 input gate. Finally, the output gate (o_t) selects features from the updated cell state c_t to
258 produce hidden state h_t , which serves as the output to the next time step. Detailed

259 equations and descriptions are provided in Supporting Information Text S2.



260
261 **Figure 4. Model architecture: (a) Encoder-decoder LSTM framework for**
262 **groundwater-level prediction. The encoder processes an input sequence of m time**
263 **steps (x_1, x_2, \dots, x_m), where each x_t contains historical groundwater levels,**
264 **precipitation, potential evaporation, and reservoir discharge. The decoder**
265 **converts the encoded temporal information into the predicted downstream**
266 **groundwater level (y_{m+1}). (b) Basic LSTM layer structure for the time step t to $t +$**
267 **1, with three essential gates including forget, input, and output gates that regulate**
268 **the cell state update and control the information flow through the network..**
269 **Abbreviations in the figure: FC, fully connected layers; GWL, groundwater level.**

270 As an important hyperparameter of encoder-decoder LSTM, temporal dependencies
271 parameter (input sequence length) significantly influences model performance,
272 particularly in capturing both short-term fluctuations and long-term trends in
273 hydrological time series (Wunsch et al., 2021). To evaluate the model performance
274 under varying input sequence length during the 2023 flood season, the dataset was
275 partitioned into a training period (January 1, 2018 to May 31, 2023) and a testing period
276 (June 1, 2023 to September 30, 2023). The methodological workflow comprised the
277 following six phases:

278 1) Data preprocessing: normalization was applied to address dimensional
279 discrepancies among heterogeneous features.

280 2) Input sequence length configuration: the length m was determined experimentally,
281 ranging from 1 to 15 days, through parametric trials. This range was chosen to
282 balance short-term and medium-term dependencies, ensuring that hydrological
283 fluctuations can be effectively captured.

284 3) Network architecture: the encoder-decoder was coupled with the LSTM
285 architecture. A fully connected layer converted decoder states into normalized
286 predictions.

287 4) Model training: the Adam optimizer was used to minimize mean squared error
288 (MSE) loss with a learning rate of 0.001. The training configuration included a batch
289 size of 32 and a maximum of 200 epochs.

290 5) Post-processing: predictions were denormalized to derive GWL values.

291 6) Performance evaluation: prediction accuracy was quantified using the coefficient
292 of determination ($R^2 \in [0,1]$) and mean squared error ($MSE \in [0, +\infty)$), with perfect
293 predictions achieved when $R^2 = 1$ and $MSE = 0$.

294 **2.3.3 Coupling Framework of Deep Learning and Multi-Objective 295 Optimization**

296 The coupled deep learning and multi-objective optimization framework was
297 developed for the Lincheng Reservoir to optimize flood control, water storage, and
298 groundwater recovery. This framework incorporates flood limited water level (FLWL)
299 variations across operational scenarios to assess the impacts of strategies on

300 multifunctional performance. The Non-dominated Sorting Genetic Algorithm II
 301 (NSGA-II) ([Deb et al., 2002](#)) was utilized to identify the Pareto frontier (i.e., the set of
 302 non-dominated solutions representing optimal trade-offs among competing objectives)
 303 under constraints.

304 **2.3.3.1 Objective Functions**

305 The developed multi-objective optimization framework incorporated the following
 306 objective functions:

307 The first is the reservoir flood risk during reservoir operation, which was quantified
 308 as the exceedance magnitude above FLWL. The corresponding objective was
 309 formulated to minimize the cumulative excessive flood damage as:

$$310 \quad \min F_1 = \sum_{t=1}^T f_1^t \quad (1)$$

311 where T represents the total number of scheduling periods, and f_1^t is the relative flood
 312 exceedance in period t , defined as:

$$313 \quad f_1^t = \begin{cases} (W_t - W_t^l) / W_t^l & W_t \geq W_t^l \\ 0 & W_t < W_t^l \end{cases} \quad (2)$$

314 where W_t and W_t^l represent the actual reservoir storage and the storage
 315 corresponding to the FLWL at the period t [L^3], respectively.

316 Water storage benefits were quantified by the ability of the reservoir to recover to the
 317 normal storage level after the flood season. The water scarcity objective aimed to
 318 minimize terminal storage deviation by:

$$319 \quad \min F_2 = \frac{W_s - W_{t=T}}{W_s} \quad (3)$$

320 where W_s denotes the storage capacity at normal storage level [L^3] and $W_{t=T}$

321 represents the reservoir storage at the end of the regulation period [L^3].

322 Groundwater recovery benefits were evaluated through post-regulation groundwater
323 level rise relative to baseline operations. The groundwater recovery objective
324 maximized groundwater level elevation gain by:

$$325 \quad \max F_3 = \frac{GWL - GWL_{base}}{GWL_{base}} \quad (4)$$

326 where GWL is the encoder-decoder LSTM predicted groundwater level under
327 operational schemes [L], and GWL_{base} denotes the baseline groundwater level at the
328 end of the flood season [L] without floodwater utilization.

329 2.3.3.2 Constraint Conditions

330 While minimizing the objective function the following constraints were used for the
331 water balance:

$$332 \quad W_t + I_{t+1} - R_{t+1} = W_{t+1} \quad (5)$$

333 where W_t and W_{t+1} are the reservoir storage volume at period t and $t+1$ [L^3]. I_{t+1}
334 and R_{t+1} are the inflow and outflow volume at period $t+1$ [L^3], respectively.

335 The constraint on ecological flow was employed as:

$$336 \quad Q_t \geq Q_t^{eco} \quad (6)$$

337 where Q_t^{eco} represents the downstream ecological flow at period t [L^3T^{-1}], calculated
338 using the Tennant method ([Tennant, 1976](#)).

339 The constraint on downstream flood safety was defined as

$$340 \quad Q_t \leq Q_{max} \quad (7)$$

341 where Q_{max} denotes the maximum discharge for downstream flood protection [L^3T^{-1}].
342 Q_{max} was determined through hydrological frequency analysis using Pearson type

343 III distribution ([Ji et al., 1984](#)).

344 Finally, a constraint on reservoir storage was applied by:

345 $W_{\min} \leq W_t \leq W_{\max}$ (8)

346 where W_{\min} and W_{\max} are the reservoir storage correspond to dead storage level and
347 normal storage level storage, respectively [L^3].

348 **2.3.3.3 Optimization Algorithm Implementation**

349 The 2023 extreme flood event in the Haihe River Basin was utilized as a case study
350 to optimize daily reservoir operations, with the outflow-to-inflow ratio as the decision
351 variable, which provides a normalized measure of reservoir release relative to incoming
352 water volumes, thereby enabling flexible and scalable optimization across varying
353 inflow conditions, while avoiding reliance on absolute outflow values that may
354 fluctuate significantly with hydrological variability.

355 To balance the competing objectives of flood risk, water scarcity, and groundwater
356 recovery using this decision variable, a multi-objective optimization model was
357 developed and solved using the Non-dominated Sorting Genetic Algorithm II (NSGA-
358 II) ([Deb et al., 2002](#)). NSGA-II enhances computational efficiency through fast non-
359 dominated sorting and crowding distance mechanisms, while exhibiting robust
360 performance across diverse engineering applications ([Verma et al., 2021](#)). Parameter
361 configurations included a population size of 100, a maximum of 1,000 generations, a
362 crossover probability of 0.9, and a mutation probability equal to the inverse of the
363 number of decision variables.

364 Iterative optimization produced constrained discharge schemes, resulting in a non-

365 dominated solution set across the three objectives defined. The projection pursuit
 366 method ([Han et al., 2025](#)) was utilized for Pareto front analysis. This method projects
 367 high-dimensional solutions onto a lower-dimensional subspace, facilitating quantitative
 368 analysis and decision-making, while mitigating the curse of dimensionality.

369 Additionally, Spearman's rank correlation coefficient was calculated for the
 370 objective function values to investigate relationships among the objectives. The
 371 Spearman's rank correlation coefficient (ρ) is calculated by:

$$372 \quad \rho = 1 - \frac{6 \sum d_i^2}{N(N^2 - 1)} \quad (9)$$

373 where d_i represents the difference between the ranks of corresponding values for each
 374 objective pair, and N denotes the number of observations.

375 **2.3.4 Evaluation of Groundwater Recharge Measures**

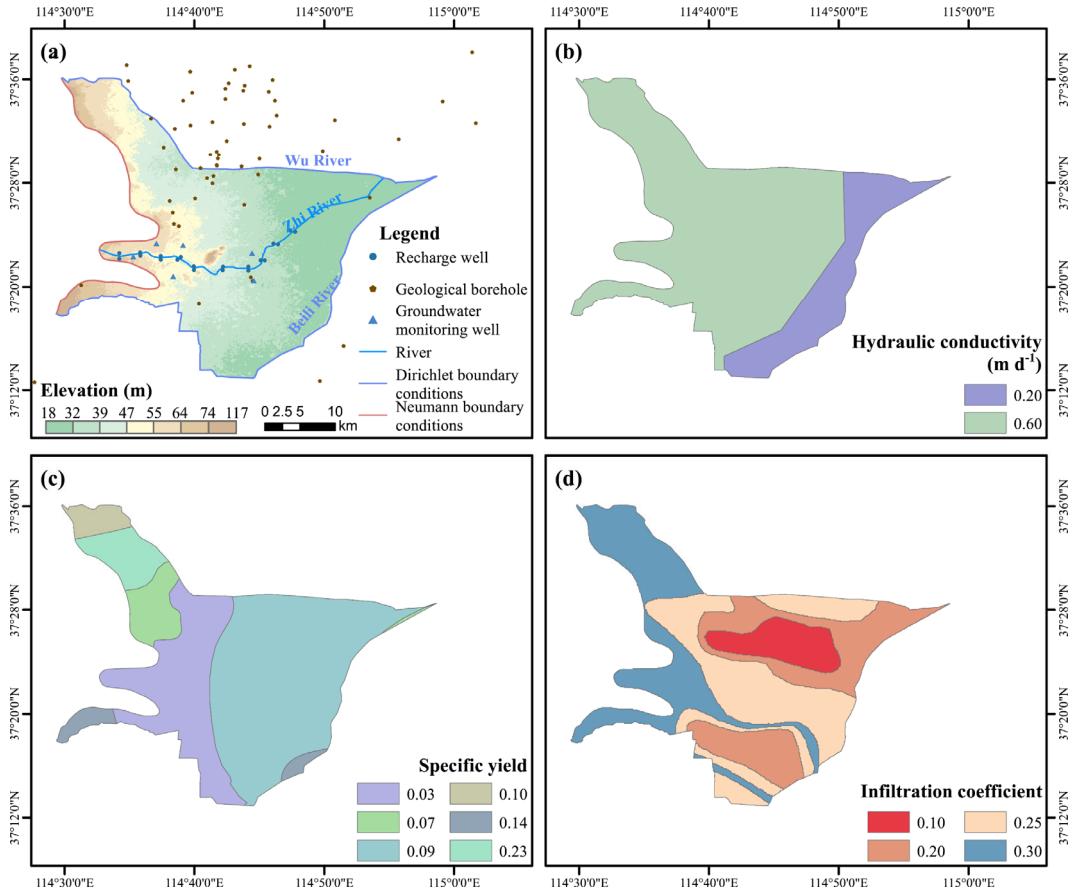
376 A three-dimensional groundwater numerical model was developed using the
 377 MODFLOW model to simulate the effectiveness of managed aquifer recharge (MAR)
 378 utilizing flood resources. Based on Darcy's law and the water balance principle, a
 379 system of partial differential equations was employed to simulate the groundwater flow
 380 numerically:

$$381 \quad \begin{cases} \mu \frac{\partial h}{\partial t} = \frac{\partial}{\partial x} (K_x \frac{\partial h}{\partial x}) + \frac{\partial}{\partial y} (K_y \frac{\partial h}{\partial y}) + \frac{\partial}{\partial z} (K_z \frac{\partial h}{\partial z}) + \varepsilon \\ h(x, y, z, t) \Big|_{t=0} = h_0 \\ h \Big|_{\Gamma_1} = h_1(x, y, z, t) \\ K_n \frac{\partial h}{\partial n} \Big|_{\Gamma_2} = q(x, y, z, t) \end{cases} \quad (10)$$

382 where K_x , K_y and K_z are the values of hydraulic conductivity [LT^{-1}] along the x , y and z
 383 coordinate axes in the simulation region Ω , h is the hydraulic head [L], ε is the

384 source/sink term [LT^{-1}], μ is the specific yield, h_0 is the initial head [L], Γ_0 is the upper
385 boundary condition, Γ_1 is the Dirichlet boundary condition, Γ_2 is the Neumann
386 boundary condition, n is the outward normal direction of the Neumann boundary, q is
387 the lateral flux per unit area and per unit time at the Neumann boundary [LT^{-1}], and K_n
388 is the hydraulic conductivity in the normal direction at the boundary [LT^{-1}].

389 The simplification of boundary conditions was critical for the groundwater numerical
390 simulation. Rivers can be simplified as Dirichlet conditions, while the boundaries
391 between different hydrogeological zones can be simplified as Neumann conditions
392 (flux boundary), with flow values calculated using Darcy's law based on multi-year
393 groundwater level data. Therefore, as shown in [Figure 5a](#), the northern Wu River and
394 the southern Beili River were designated as Dirichlet boundary conditions. The western
395 boundary, representing the boundary between the Taihang Mountains and the North
396 China Plain, was designated as a Neumann boundary condition. The main channel of
397 Zhi River was implemented as an internal river boundary condition.



398

399 **Figure 5. Overview of the 3D groundwater modeling area in the Zhi River basin**
400 for (a) boundary conditions and key features, including recharge wells (blue dots),
401 geological boreholes (black dots), monitoring wells (blue triangle), with elevation
402 contours in meters, and distributions of (b) hydraulic conductivity, (c) specific
403 yield, and (d) infiltration coefficient.

404 Based on borehole data (Figure 5a), the aquifer system was vertically discretized into
405 three layers: an unconfined aquifer (Layer 1), a semi-confined aquifer (Layer 2), and a
406 confined aquifer (Layer 3). The absence of consistent impermeable layers between the
407 unconfined and semi-confined aquifers results in a strong hydraulic connection. Given
408 that the study area is located in a typical shallow groundwater depression cone (Figure
409 2b), Layers 1 and 2, with depths ranging from 13 to 60 m, and 80 to 150 m, respectively,
410 were selected as the primary focus of this study.

411 Source terms included lateral recharge, precipitation infiltration, and irrigation return

412 flow. Sink terms included lateral discharge, groundwater extraction, and actual
413 evapotranspiration. However, the groundwater table in this region typically exceeds 4
414 m in depth, which surpasses the maximum evaporation depth for unconfined aquifers.
415 When the groundwater table exceeds this maximum depth, the connection between
416 groundwater and evapotranspiration weakens or disappears ([Condon and Maxwell,](#)
417 [2019](#)). Consequently, actual evapotranspiration was set to zero in the sink terms.
418 Detailed calculations of the source and sink terms are provided in Supporting
419 Information Text S3. To assess the impact of the omission of evapotranspiration and the
420 flow boundary condition, a sensitivity analysis was conducted. The results of this
421 analysis are presented in Supporting Information Text S4.

422 The groundwater level on May 31, 2023 was set as the initial hydraulic head, with
423 daily stress periods spanning June 1, 2023 to September 30, 2023 (122 simulation days).
424 To evaluate MAR effectiveness, this study established 20 recharge wells along a cross-
425 section of the Zhi River (see [Figure 5a](#)). Six recharge intensity gradients (50–300 $\text{m}^3 \text{d}^{-1}$
426 with increments of $50 \text{ m}^3 \text{ d}^{-1}$) were implemented to quantify the effects of MAR using
427 floodwater.

428 **3. Results**

429 This section presents the outcomes derived from the integrated multi-scenario
430 framework coupling deep learning with multi-objective optimization and the three-
431 dimensional groundwater numerical model. We assessed the performance of the
432 encoder-decoder LSTM model in predicting groundwater levels and the groundwater
433 numerical model in predicting groundwater dynamics. Furthermore, we characterize

434 the Pareto frontier distributions, identify the best and worst solutions for each objective
435 function across the different scenarios and quantify the effects of MAR under varying
436 floodwater rates.

437 **3.1 Ecological Flow and Maximum Discharge**

438 As critical constraints, the determination of ecological flow and maximum discharge
439 directly influenced the optimization outcomes. The Tennant method was employed to
440 calculate historical monthly average river flow, which served as the baseline for
441 ecological flow assessment due to its wide applicability under limited flow data
442 conditions. Long-term records showed recurrent zero-discharge episodes during the
443 flood season, which depressed the multi-year monthly means. As an illustration, during
444 the 2023 flood season there were 57 consecutive zero-flow days from June 1 to July 27
445 (Figure S1 in Supporting Information). Accordingly, to maintain suitable aquatic
446 habitats, ecological flow thresholds were set at 60% of monthly averages during flood
447 seasons and 30% in non-flood seasons ([Table 1](#)), since flows within 60–100% of the
448 natural regime sustain good habitat conditions, while 30–60% meet basic ecological
449 requirements ([Tennant, 1976](#)). Implementing these ecological flow constraints in 2023
450 eliminated downstream river drying during the flood season, reducing zero-flow days
451 from 57 to 0, which effectively restored continuous flow connectivity and improved the
452 stability of downstream aquatic habitats.

453 **Table 1. Results of multi-year average ecological river flow calculations (1961–
454 2023).**

Month	Average River Flow ($\text{m}^3 \text{s}^{-1}$)	Ecological River Flow ($\text{m}^3 \text{s}^{-1}$)
-------	--	---

1	0.113	0.034
2	0.105	0.032
3	0.574	0.172
4	0.871	0.261
5	1.075	0.322
6	0.833	0.500
7	1.551	0.930
8	2.339	1.404
9	0.584	0.350
10	0.723	0.217
11	0.474	0.142
12	0.294	0.088

455 Historical extreme discharge events were analyzed through empirical frequency
 456 analysis of annual maximum discharge records. Documented events included a 200-
 457 year return period flood ($2,448 \text{ m}^3 \text{ s}^{-1}$) in 1963 and a 100-year return period flood (1,016
 458 $\text{m}^3 \text{ s}^{-1}$) in 1996. The unified empirical frequency analysis, combined with Pearson Type
 459 III distribution curve fitting, demonstrated excellent model performance ($R^2=0.96$),
 460 confirming the ability of the distribution to statistically characterize discharge extremes
 461 (Figure S2 in Supporting Information). When adopting a 2% exceedance probability
 462 standard (i.e., a 50-year return period), the maximum discharge for flood safety was
 463 determined to be $734 \text{ m}^3 \text{ s}^{-1}$, which was selected to ensure adequate flood protection for
 464 downstream urban areas.

465 These two constraints, including minimum ecological flow and maximum discharge
 466 threshold, jointly defined the feasible solution space of the optimization process. The
 467 minimum ecological flow limits excessive water retention in the reservoir, thereby
 468 ensuring continuous downstream flow connectivity even under low-inflow conditions.
 469 In contrast, the maximum discharge threshold limits excessive flood releases, thereby
 470 preventing downstream flood hazards. Consequently, these constraints confined the

471 optimization results within a reasonable range, ensuring that downstream ecological
472 requirements were satisfied while maintaining sufficient flood-control capacity.

473 **3.2 Groundwater Level Prediction Effect**

474 To optimize the predictive accuracy of the encoder-decoder LSTM model, we
475 evaluated its performance by tuning the input sequence length (m), which governs the
476 temporal dependency window through which the model captures hydrological memory.

477 In this context, m represents the time horizon over which antecedent groundwater levels,
478 precipitation, potential evaporation, and reservoir discharge collectively influence
479 groundwater dynamics. Short sequences may fail to capture delayed hydrological
480 feedback, whereas long sequences may introduce redundant temporal information.

481 Comparative performance metrics across different m varying from 1–15 days are
482 illustrated in [Figures 6a–o](#). While maintaining superior training performance ($R^2 > 0.93$,

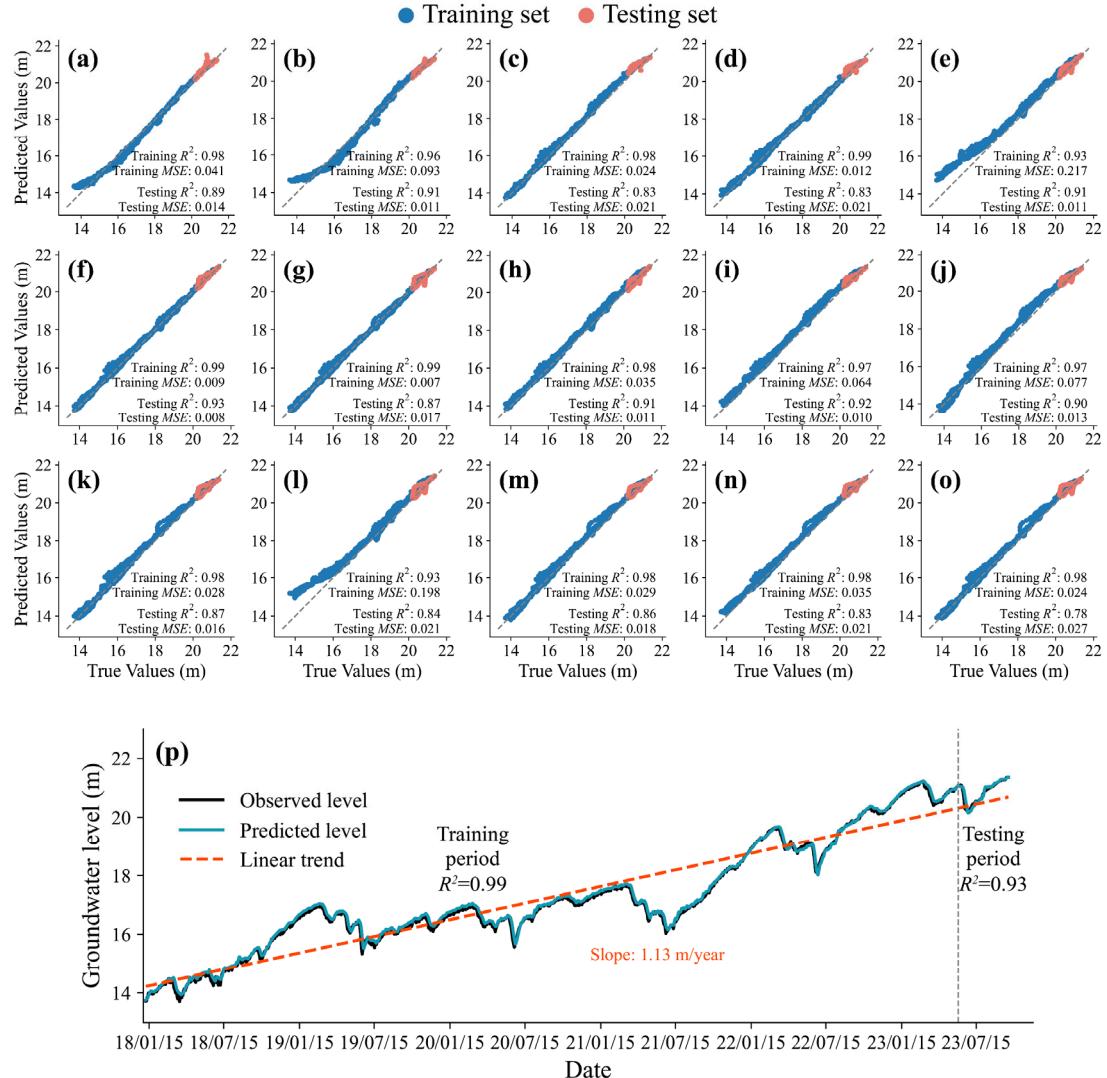
483 $MSE < 0.2$) for all m , testing accuracy showed notable variability. m of 5–11 days
484 yielded robust testing set performance ($R^2 > 0.85$, $MSE < 0.02$), indicating strong

485 correlations between hydrological inputs (including precipitation, evaporation, and
486 discharge) and groundwater levels. The model performed optimally on the testing set
487 at $m = 6$ days ($R^2 = 0.93$, $MSE = 0.008$), followed closely by $m = 9$ days ($R^2 = 0.92$,

488 $MSE = 0.010$). This superior performance corresponds to the characteristic lag between
489 precipitation, evaporation, reservoir releases, and groundwater level responses,

490 indicating that the model effectively captured short-term surface–subsurface
491 interactions. Accordingly, these two configurations were selected for integration into
492 the coupling framework in subsequent studies: $m = 6$ for the primary groundwater level

493 predictions in the coupling framework due to its superior performance, and $m = 9$ for
 494 subsequent sensitivity analyses to assess the robustness of optimization outcomes.



496
 497 **Figure 6. Results of the groundwater level prediction of the encoder-decoder**
 498 **LSTM Model: (a)-(o) model performance on the training and testing sets for input**
 499 **sequence length of 1-15 days; (p) fitting effect of groundwater dynamics with an**
 500 **input sequence length of 6 days, where the black solid line, blue solid line, and red**
 501 **dashed line represents the observed, predicted groundwater level, and the trend of**
 502 **groundwater level changes during the prediction period.**

503 The encoder-decoder LSTM model performed effectively in predicting groundwater
 504 dynamics at an input sequence length of 6 days. As demonstrated in [Figure 6p](#), the
 505 model achieved high predictive performance, with training period accuracy ($R^2=0.99$)

506 and testing period generalization capacity ($R^2=0.93$).

507 The predicted groundwater level series closely matched the observed values over
508 time throughout the monitoring period, even during extreme water level fluctuations.

509 Additionally, groundwater levels exhibited an upward trend throughout the prediction
510 period, with a long-term trend (see the fitted line in Figure 6p) indicating an average
511 annual groundwater recovery rate of 1.13 m, which resulted from the implementation
512 of effective groundwater management and control measures in the study area.
513 Consequently, this $m = 6$ configuration was selected for groundwater prediction in
514 subsequent multi-objective optimization.

515 **3.3 Pareto Frontier Distribution**

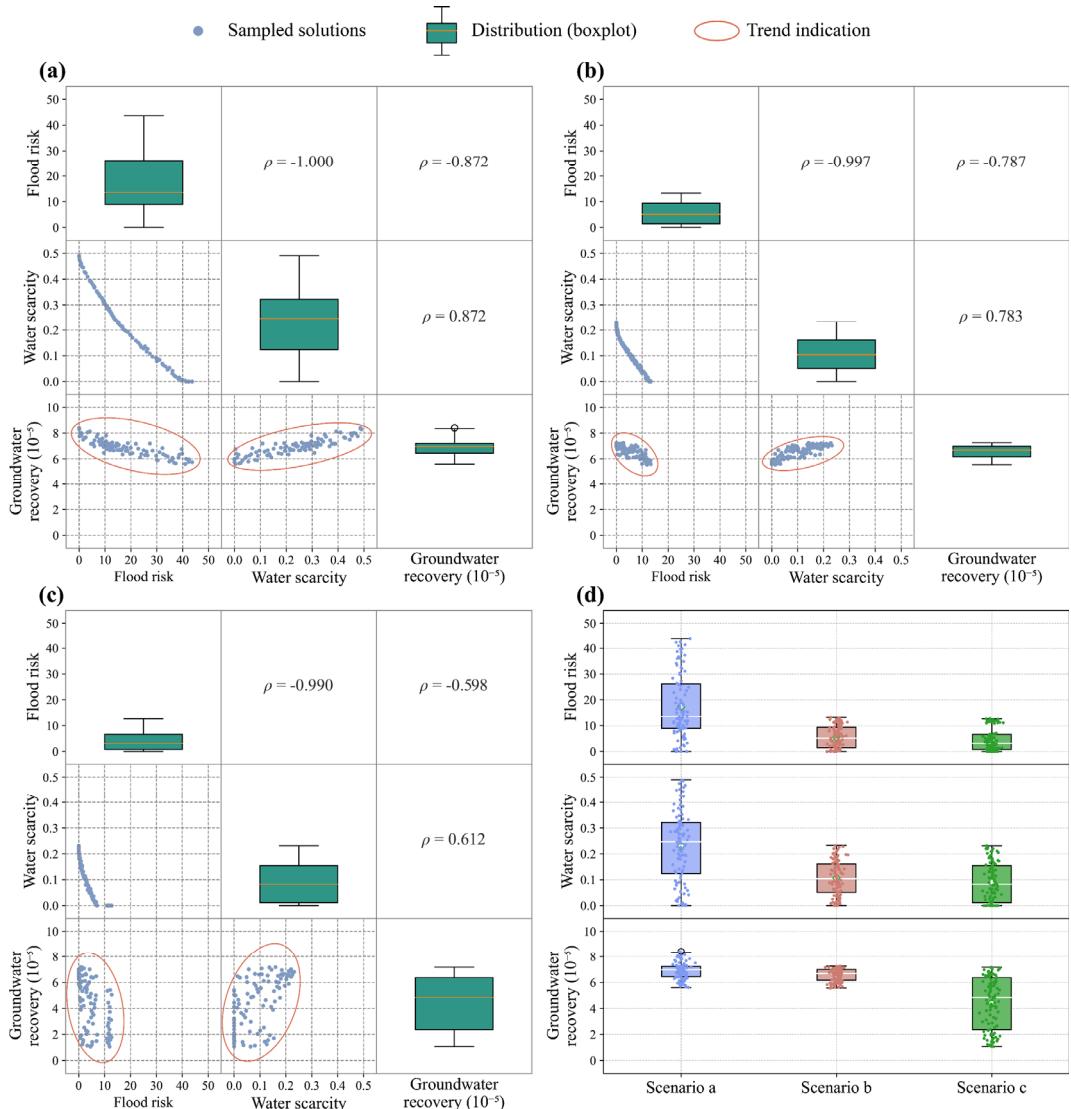
516 Based on the previously determined input length of $m = 6$ days, the encoder-decoder
517 LSTM model was incorporated into the coupled deep learning and multi-objective
518 optimization framework. Through iterative optimization computations, this study
519 generated the two-dimensional projections of the original three-objective Pareto front
520 for the Lincheng Reservoir system under varying FLWL scenarios ([Figure 7](#)). The
521 depicted box plots (the diagonal graphs in [Figure 7](#)) quantitatively characterize the
522 distributional properties of each objective, while the scatter plots (the off-diagonal
523 graphs in [Figure 7](#)) illustrate the inverse and positive relationships among competing
524 objectives. Additionally, the Spearman's rank correlation coefficient (ρ) reflected the
525 strength of these relationships. Negative correlations ($\rho < 0$) denote inverse
526 relationships, where improvements in one objective (e.g., reducing reservoir flood risk)
527 come at the expense of another (e.g., increasing reservoir water scarcity), necessitating

528 compromises in decision-making to balance competing priorities. In contrast, positive
529 correlations ($\rho > 0$) indicate advancements in one objective simultaneously support or
530 enhance the other, enabling optimization without inherent conflicts. [Figure 7](#) shows that
531 there exists a detectable inverse relationship between the reservoir flood risk and
532 reservoir water scarcity objectives across all scenarios. The negative ρ values among
533 these two factors (all below -0.99) in all scenarios indicated that it was challenging to
534 reduce both reservoir flood risk and water scarcity simultaneously through optimized
535 scheduling. Reservoir water scarcity objectives exhibited positive relationships with
536 groundwater recovery targets ($\rho > 0.60$ in all scenarios), while reservoir flood risk
537 objectives demonstrated inverse characteristics with groundwater recovery ($\rho < -0.59$
538 in all scenarios). When reservoir discharge decreased, water storage in the reservoir
539 increased, more inflow was retained within the reservoir, and consequently elevating
540 reservoir flood risk during the flood season while reducing reservoir water scarcity risk
541 at the end of the flood season. Meanwhile, the reduced downstream release decreased
542 the available infiltration, leading to weaker groundwater recovery effects. Therefore,
543 reservoir flood risk exhibited inverse relationships with both reservoir water scarcity
544 and downstream groundwater recovery, whereas reservoir water scarcity and
545 groundwater recovery showed a positive relationship. This outcome is consistent with
546 the fundamental principles of water balance in the hydrological cycle.

547 The groundwater recovery targets calculated in all scenarios were relatively small,
548 with values on the order of 10^{-5} . These values are dimensionless, as they were calculated
549 as the ratio of groundwater recovery to the original groundwater level, representing

550 relative rather than absolute changes. This confirmed that relying solely on discharge
551 flow regulation for groundwater recovery was substantially limited and further
552 indicated that, within the imposed operational constraints, the variations in reservoir
553 discharge induce only minor changes in downstream river flow compared with other
554 groundwater source and sink terms.

555 With the transition from Scenario *a* to Scenario *c*, ρ between reservoir water scarcity
556 and groundwater recovery decreased from 0.872 to 0.612, while that between reservoir
557 flood risk and groundwater recovery was also weakened from -0.872 to -0.598. This
558 weakening of both positive and negative correlations arises because raising the FLWL
559 allowed more water to be stored in the reservoir, increasing its storage capacity. As a
560 result, both reservoir flood risk and reservoir water scarcity risk declined, whereas
561 groundwater recovery diminished because reduced downstream releases recharge led
562 to less infiltration. This adjustment increased the influence of storage capacity on both
563 flood risk and reservoir water scarcity, while groundwater recovery remained primarily
564 controlled by reservoir discharge. Consequently, elevating the FLWL across different
565 scenarios reduces the direct competition between flood control and reservoir water
566 storage, enabling simultaneous enhancements in flood control safety and water storage
567 benefits. These shifts altered the mechanisms linking the objectives, thereby attenuating
568 the strength of their inverse or positive relationships.



569

570 **Figure 7. Two-dimensional matrix diagram of the Pareto front for an input**
 571 **sequence length of 6 days for (a) Scenario a, (b) Scenario b, and (c) Scenario c,**
 572 **illustrating the relationships among the three optimization objectives: flood risk,**
 573 **water scarcity, and groundwater recovery. Blue scatter points represent the**
 574 **optimized solutions, red ellipses indicate the overall inverse or positive trends**
 575 **between objectives, and ρ quantifies the strength and direction of these**
 576 **relationships. Green boxplots along the diagonal show the distribution of each**
 577 **objective within its respective scenario. (d) compares the distributions of the three**
 578 **objectives across the three FLWL scenarios, revealing the sensitivity of the**
 579 **optimization outcomes to different reservoir water-level constraints.**

580 Adjustments to FLWL induced substantial changes in the Pareto front objective
 581 values (Figure 7d). Increasing FLWL (Scenarios a-c) reduced flood risks by 84.9% and
 582 water scarcity risks by 61.9%, confirming that an increase in FLWL within safe

583 operational thresholds enables the dual-objective optimization of flood control and
584 water storage. This improvement arises because a higher FLWL increases the storage
585 capacity of the reservoir, thereby lowering the probability of FLWL exceedance while
586 allowing more water to be retained for subsequent use, which consequently reduces
587 water scarcity losses at the end of the flood season. However, the accompanying
588 decrease in downstream discharge reduces river-aquifer interactions, thereby
589 weakening groundwater recovery by 22.2%. Groundwater recovery targets remained
590 positive (>0) across all scenarios, demonstrating that maintaining regulated ecological
591 flow thresholds effectively mitigates groundwater level deterioration even under
592 diminished total discharge conditions. However, groundwater recovery remained
593 limited across all scenarios. This suggested that while maintaining ecological flow,
594 which facilitated groundwater recovery, additional effective measures were still
595 required to achieve rapid groundwater recovery.

596 **3.4 Best and Worst Solution for Each Objective**

597 As demonstrated in previous sections, distinct positive or inverse relationships
598 existed between different objectives. In the next step, the best and worst solutions for
599 each objective were then analyzed to assist decision-makers in selecting management
600 strategies based on different priorities. [Figure 8](#) illustrates the solutions for each
601 objective across scenarios. The results revealed substantially larger discharge volumes
602 during peak flood periods (July 30 to August 1) compared to other periods under all
603 scenario-solution combinations.

604 The best solution for flood control (the first row graphs in [Figure 8](#)) kept the reservoir

605 levels below the corresponding FLWL throughout all operational periods (F_1 in
606 Equation 3 equals 0). Given the pronounced inverse relationships between flood risk
607 and reservoir water scarcity objectives, this solution also represented the worst solution
608 for water storage in Scenarios a and b ([Figures 8a4](#) and [8b4](#)) or closely approximates it
609 in Scenario c ([Figure 8c4](#)), thereby highlighting the direct conflict between minimizing
610 reservoir flood risk throughout the flood season and ensuring adequate storage at the
611 end of the flood season. The stringent implementation of FLWL operation strategies
612 resulted in waste of floodwater resources. The floodwater utilization rate (i.e., utilized
613 flood volume/total flood volume) stayed below 25% across all scenarios, even dropping
614 below 5% in Scenario *a*. This outcome arose because higher outflows led to lower
615 reservoir levels, preventing full utilization of storage capacity. As a result, floodwater
616 was primarily released rather than being retained for future use, resulting in low
617 floodwater utilization efficiency. Consequently, reservoir water levels at the end of the
618 flood season became critically low, making rapid restoration to normal storage level
619 operationally challenging.

620 The best solution for water storage (the third row graphs in [Figure 8](#)) improved
621 floodwater utilization, with resource utilization rates approaching 40% across all
622 scenarios. This operational strategy raised reservoir levels to normal storage capacity
623 by the end of the flood season ($F_2 = 0$). However, this solution also represented the
624 worst solution for flood control (the second row graphs in [Figure 8](#)). This
625 implementation resulted in reservoir level exceedances above FLWL for 62 days
626 (Scenario *a*), 51 days (Scenario *b*), and 26 days (Scenario *c*), respectively. This trend

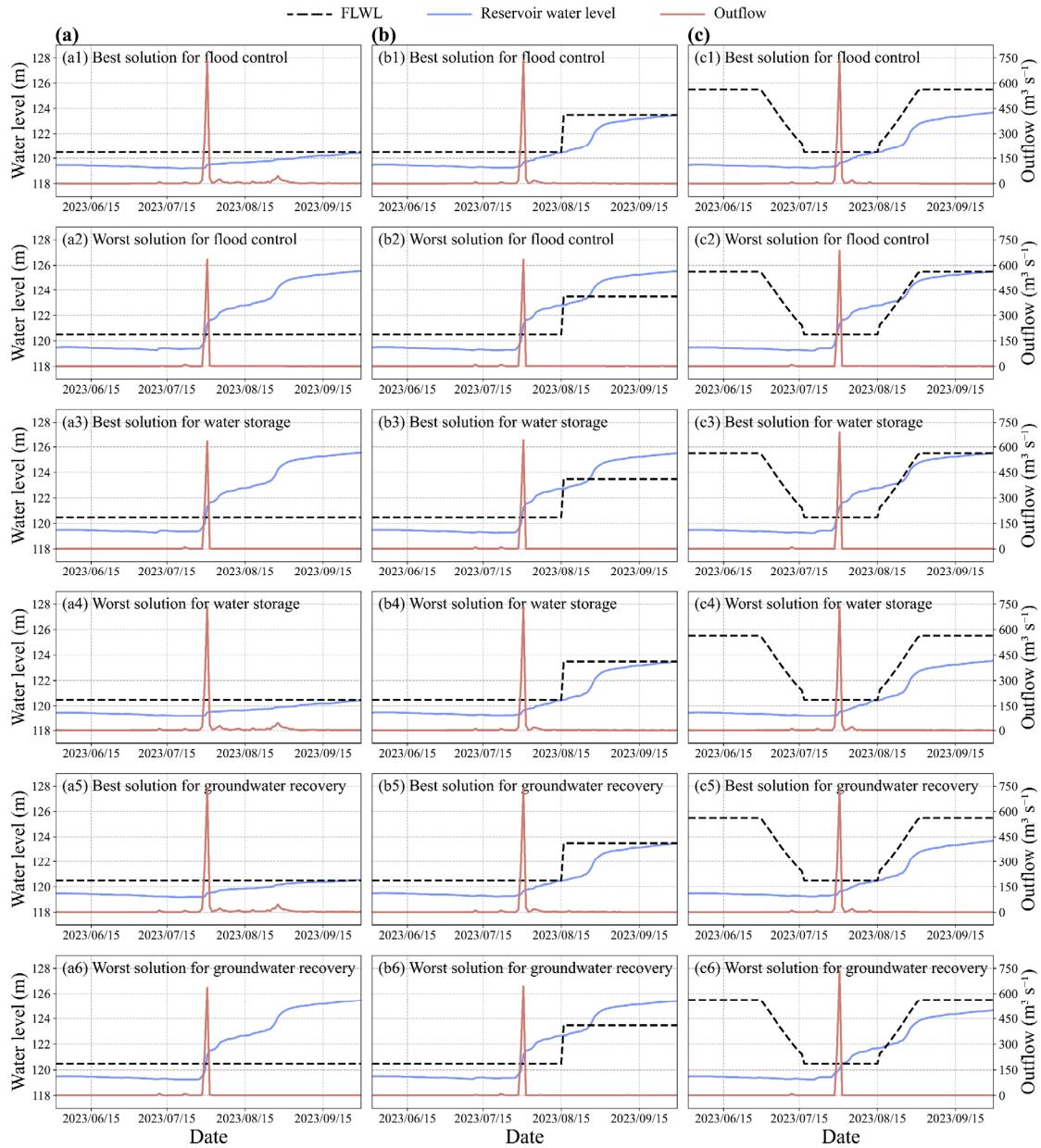
627 was driven by the increase in FLWL, which expanded the allowable storage space,
628 thereby reducing the frequency of exceedances under the same storage strategy.

629 The best solution for groundwater recovery (the fifth row graphs in [Figure 8](#)) within
630 the optimization framework showed operational patterns similar to the best solution for
631 flood control (the first row graphs in [Figure 8](#)), with reservoir levels remaining below
632 the FLWL. While this strategy maximized the groundwater recovery objective relative
633 to other feasible operations, its absolute efficacy remained limited, since a considerable
634 portion of floodwater releases could not be effectively converted into subsurface
635 storage. This limitation arose because the duration of flood peaks was short, and the
636 limited short-term infiltration restricted substantial groundwater recharge during the
637 flood season. This also highlights the limitation of relying solely on reservoir release
638 regulation for groundwater recovery.

639 The worst solution for groundwater recovery (the sixth row graphs in [Figure 8](#)) was
640 characterized by minimal reservoir discharge, which substantially reduced groundwater
641 recharge and produced the lowest recovery effects. Because of the reduced outflow, the
642 reservoir retained higher water levels by the end of the flood season, generally
643 remaining close to the normal storage level.

644 While the best and worst solutions for each objective cannot achieve optimization of
645 all objectives simultaneously, a moderate elevation of FLWL facilitated simultaneous
646 reduction of both flood and water scarcity risks. However, as indicated by the limited
647 groundwater recovery across scenarios, additional measures such as managed aquifer
648 recharge (MAR) are necessary to enhance groundwater recovery. Accordingly, the

649 following section evaluates groundwater numerical simulations and quantifies the
 650 impacts of MAR under varying recharge rates.

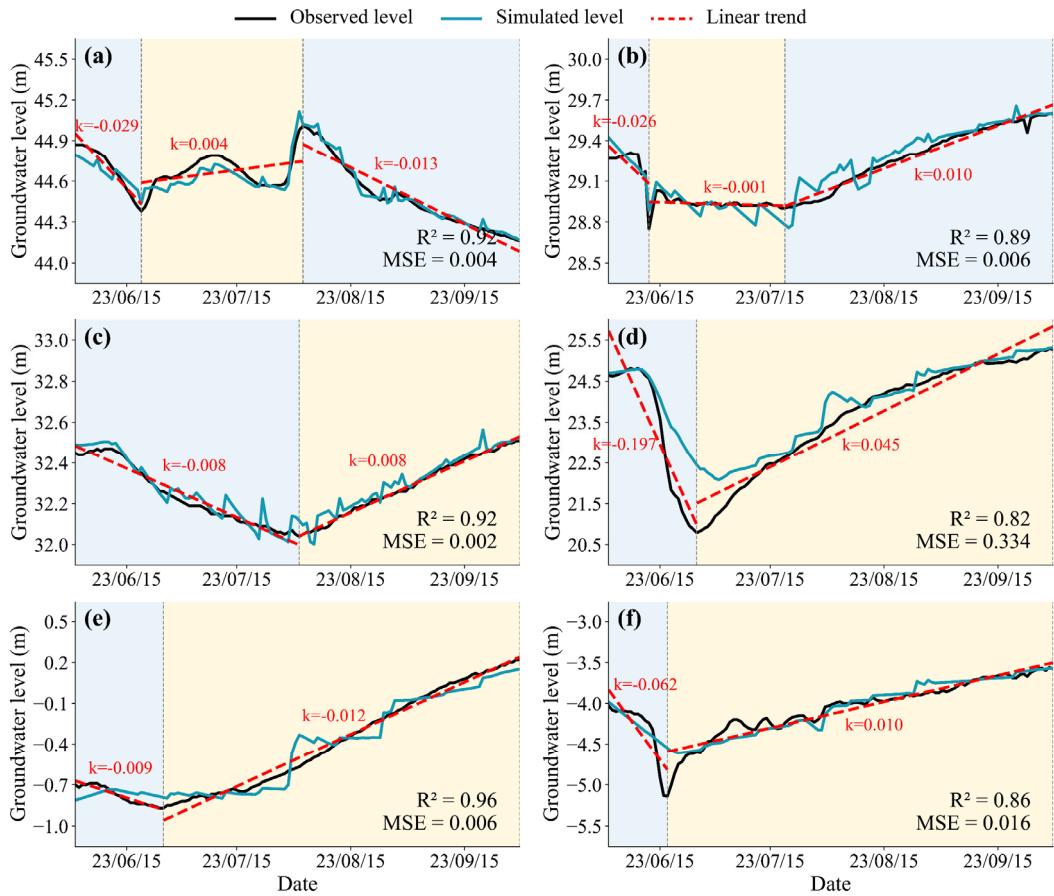


651
 652 **Figure 8. Best and worst solution for each objective under Scenarios *a*, *b*, and *c*.**
 653 The first and second row graphs show the best and worst solutions for flood
 654 control, the third and fourth row graphs for water storage, and the fifth and sixth
 655 row graphs for groundwater recovery. The best solution for flood control
 656 maintains reservoir levels below corresponding FLWL throughout all operational
 657 periods, while the best solution for water storage achieves normal storage level by
 658 the end of flood season.

659 **3.5 Effectiveness of Groundwater Numerical Simulation**

660 To further investigate the effective floodwater utilization strategies, the proposed
661 MAR scheme was tested, whereby accurate simulation of groundwater dynamics
662 served as the essential basis for evaluating the feasibility of this strategy. To evaluate
663 MAR feasibility, the MODFLOW model was employed to simulate groundwater
664 dynamics. A comparative analysis of simulated and observed groundwater levels at six
665 monitoring wells (i.e., Zhongzhang, Xiyin, Tunli, Beicun, Maoshanying, and Longyao)
666 is presented in [Figure 9](#).

667 Overall, the simulated results demonstrated strong agreement with observed trends,
668 indicating that the model effectively captured groundwater level dynamics. Most
669 monitoring wells exhibited satisfactory calibration performance, with R^2 ranging from
670 0.82 to 0.96 and MSE between 0.0006 and 0.016 m, confirming the robustness of the
671 simulation accuracy. These results validated the applicability of MODFLOW for
672 providing reliable support for the quantitative assessment of managed aquifer recharge
673 effectiveness.



674

675 **Figure 9. Comparison of measured and simulated groundwater levels at different**
 676 **monitoring wells. (a) Zhongzhang, (b) Beicun, (c) Xiyin, (d) Tunli, (e)**
 677 **Maoshanying, and (f) Longyao, with the black solid line representing the**
 678 **measured level, the blue solid line the simulated level, and the red dashed line**
 679 **the long-term linear trend. Different background colors distinguish various trend**
 680 **stages during the simulation period.**

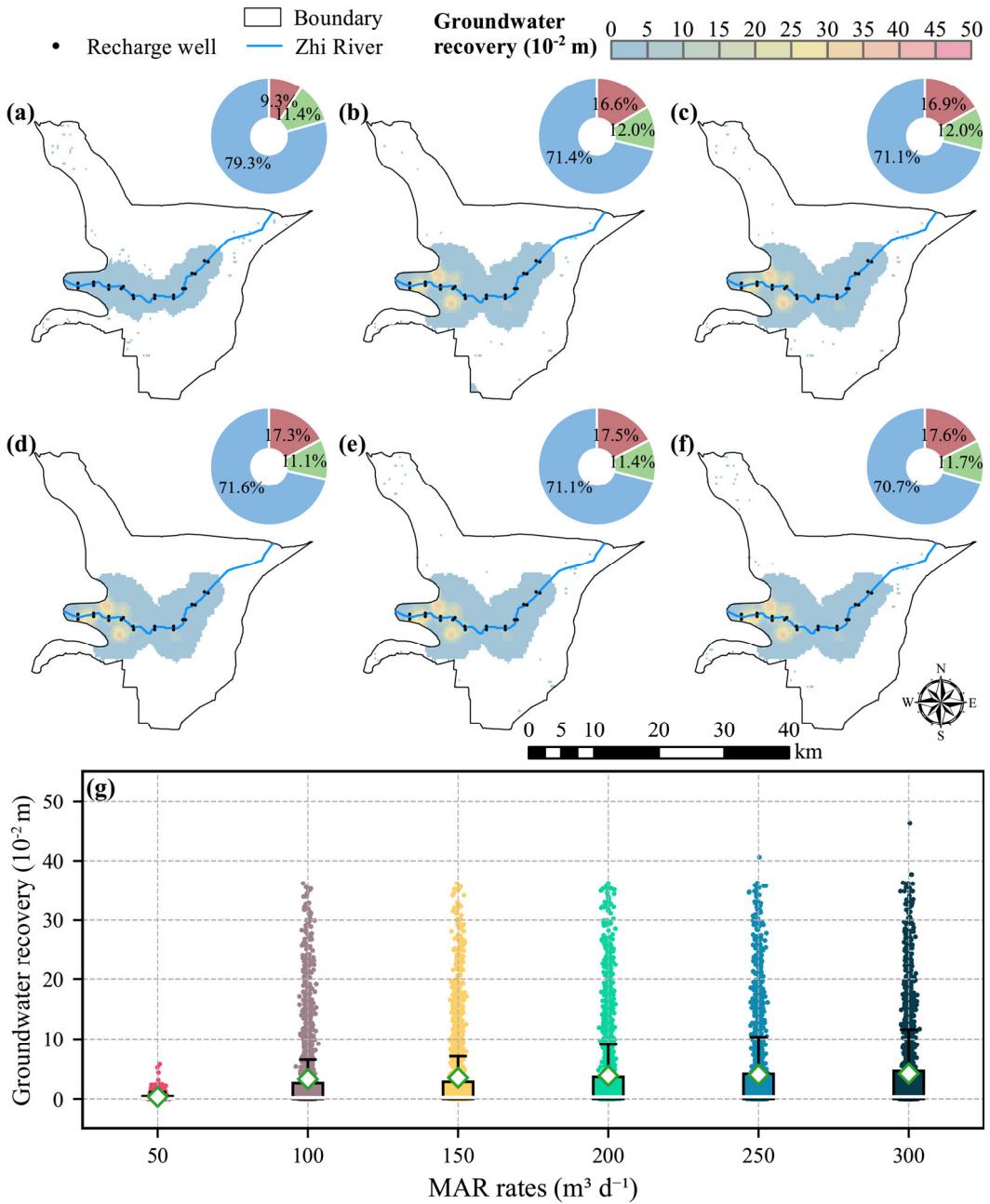
681 During the simulation period, the groundwater levels of six monitoring wells
 682 exhibited distinct stage trends. For most wells (excluding Zhongzhang and Beicun in
 683 Figures 9a-b), the levels exhibited a decline in groundwater levels before recovering.
 684 In contrast, the measured and simulated groundwater level at Zhongzhang well
 685 exhibited a declining trend, followed by a slow increase, and a followed decline. The
 686 level at Beicun (Figure 9b) also exhibited a trend of decline, whereby the second phase
 687 was characterized by stabilization, and a followed increase in groundwater levels. The
 688 initial decline in groundwater levels across nearly all wells can be associated with high

689 groundwater extraction used for spring irrigation, while the subsequent increases were
690 driven by substantial precipitation during the flood season. These deviations can be
691 attributed to localized groundwater extraction, which induced additional fluctuations in
692 the groundwater levels at these two wells.

693 **3.6 Investigation on Efficient Utilization of Flood Resources**

694 Leveraging the validated MODFLOW model, we quantified the spatial and
695 quantitative effects of MAR using floodwater resources under varying recharge rates,
696 as illustrated in [Figure 10](#). Due to varying hydrogeological conditions in the area, the
697 groundwater recovery levels exhibited spatial heterogeneity. As can be seen,
698 groundwater recovery primarily occurred around the Zhi River, with more pronounced
699 recovery observed in the western regions compared to the eastern regions. This
700 difference might arise from the smaller specific yield in the west, where limited aquifer
701 storage capacity causes the same recharge volume to produce a high rise in groundwater
702 levels. By assuming recharge intensities from the 20 recharge wells ranging from 50 to
703 $300 \text{ m}^3 \text{ d}^{-1}$, the proportion of the effective groundwater recovery area also varied and
704 increased from 9.3 to 17.6%. Simultaneously, the average recovery in the effective
705 recovery areas increased from 0.01 m to 0.09 m. The maximum groundwater recovery
706 exhibited a distinct three-phase nonlinear pattern in response to varying recharge rates.
707 In phase I, with recharge rates of $50 - 100 \text{ m}^3 \text{ d}^{-1}$, the maximum recovery level surged
708 from 0.06 m to 0.36 m, as the aquifer's recharge potential was high at this stage. In
709 phase II, with recharge rates of $100-200 \text{ m}^3 \text{ d}^{-1}$, it stabilized at 0.36 m, suggesting a
710 diminishing response as additional recharge no longer produced proportional increases

711 in groundwater levels. In phase III, with the highest recharge rates of 200-300 m³ d⁻¹,
712 the recovery gradually increased from 0.36 m to 0.46 m, since enhanced recharge
713 expanded the groundwater recovery zone, leading to a further rise in water levels but
714 with a smaller magnitude than that in phase I. As shown in Figure 10g, the recovery
715 rate significantly improved between recharge rates of 50 and 100 m³ d⁻¹ ($p < 0.05$),
716 indicating a marked response within this recharge rates range. This pattern demonstrates
717 that increasing recharge intensity enhances groundwater recovery levels, but the effect
718 might be weakened as the aquifer approaches its limits.



719
720 **Figure 10. Spatial distribution of groundwater recovery under varying recharge**
721 **rates in the 20 wells. (a) 50, (b) 100, (c) 150, (d) 200, (e) 250, and (f) 300 $\text{m}^3 \text{d}^{-1}$. Blue**
722 **lines represent river, and black dots denote recharge wells. The inset pie chart**
723 **illustrated the proportion of areas with different recovery levels: blue indicated**
724 **unrecovered areas (recovery = 0 m), green showed inefficient recovery areas ($0 <$**
725 **recovery $< 0.01 \text{ m}$), and red denoted effective recovery areas (recovery $> 0.01 \text{ m}$).**
726 **(g) the distributions of recovery values ($> 0 \text{ m}$) across different recharge rates,**
727 **presented as boxplots and scatter points, where mean values are denoted by**
728 **diamonds.**

729 This pattern is consistent with the findings of Samanta et al. (2020) on recharge
730 volume-dependent infiltration rate thresholds. Therefore, the effective implementation

731 of MAR should consider the spatial heterogeneity of aquifer permeability, optimize
732 recharge rates, and consider land use constraints ([Owuor et al., 2016](#)) to avoid
733 inefficient percolation zones, while balancing recovery efficacy with engineering costs.

734 **4. Discussion**

735 While the aforementioned results demonstrate the effectiveness of the proposed
736 floodwater utilization framework, including the performance evaluation of the
737 groundwater numerical model, and the quantification of MAR effects using floodwater
738 resources, several key aspects merit further examination to contextualize these findings
739 and inform future applications. Accordingly, this discussion first examines the
740 sensitivity of optimization outcomes to variations in the encoder-decoder LSTM input
741 sequence length, MAR recharge rates and decision variable, before addressing the
742 limitations and potential risks.

743 **4.1 Impact of Input Sequence Length of the Groundwater Level Prediction**

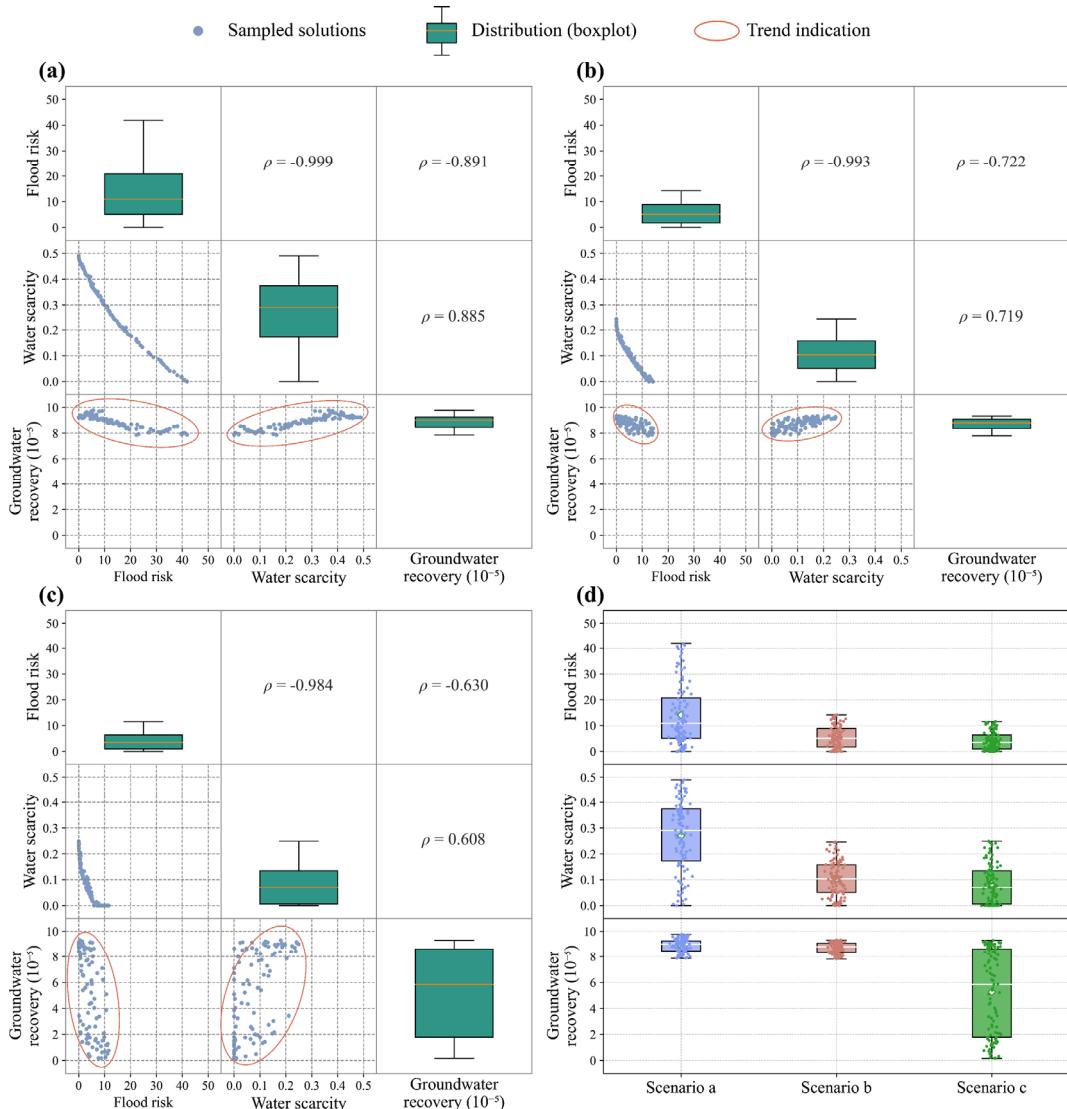
744 **Model on Optimization Results**

745 The developed deep learning and multi-objective optimization framework effectively
746 quantified competing objectives. However, as one of the key hyperparameters, which
747 are user-defined settings that govern model architecture and training, the input sequence
748 length m has a significant impact on the predictive performance of LSTM models
749 ([Gauch et al., 2021](#); [Hosseini et al., 2024](#)). While the configuration with $m = 6$ days was
750 selected as suitable based on comprehensive training and testing performance metrics,
751 the model with $m = 9$ days showed comparable robustness ([Figure 6](#)). To examine the

752 impact of this hyperparameter on optimization results, the 9-day model was integrated
753 into the multi-objective framework, generating alternative Pareto front solutions
754 ([Figure 11](#)).

755 The results indicated that adjustments to m affect the objective function values across
756 all objectives. Flood risks and water scarcity showed no significant differences ($p >$
757 0.05) between the two configurations analyzed. Groundwater recovery objectives
758 exhibited significant differences ($p < 0.05$) but maintained consistently positive
759 objective function values. This demonstrated that the continuous maintenance of river
760 flow was critical for effective groundwater recovery.

761 Notably, changes of input sequence length (from 6 to 9 days) did not alter inter-
762 objective inverse or positive effects. Flood risk maintained inverse relationships with
763 reservoir water scarcity and groundwater recovery, while water scarcity and
764 groundwater recovery retained a positive relationship. Similarly, this change of m also
765 did not affect the response of objectives across the different scenarios. When FLWL
766 was elevated, flood risks and reservoir water scarcity losses significantly decreased (p
767 < 0.05) when moving from $m = 6$ to 9 days configuration. At the same time, the inverse
768 or positive relationships between objectives became weakened (the absolute value of ρ
769 decreased). This indicated that increasing FLWL could reduce competition between
770 multi-objectives optimization and promote system balance. Changes in m did not alter
771 the relationships between objectives. However, its limitations still persist, as variations
772 in sequence length may impact model performance under different hydrological
773 conditions or when applied to new scenarios.



774

775 **Figure 11. Two-dimensional matrix diagram of Pareto front competition when**
776 **input sequence length was set to 9 days under (a) Scenario a, (b) Scenario b, and**
777 **(c) Scenario c, illustrating the relationships among the three optimization**
778 **objectives: flood risk, water scarcity, and groundwater recovery. The blue scatters**
779 **show relationships between objectives, the red ellipses indicate the overall trend**
780 **of the scatter points, the green box show the distribution of each objective within**
781 **its respective scenario., and ρ reflects the strength of positive or inverse**
782 **relationships. (d) compares the distributions of the three objectives across the**
783 **three FLWL scenarios, revealing the sensitivity of the optimization outcomes to**
784 **different reservoir water-level constraints.**

785

4.2 Impact of Managed Aquifer Recharge on Optimization Results

786

Based on the quantified benefits of MAR on groundwater recovery as demonstrated

787

in previous sections, this section explores its broader implications for the multi-

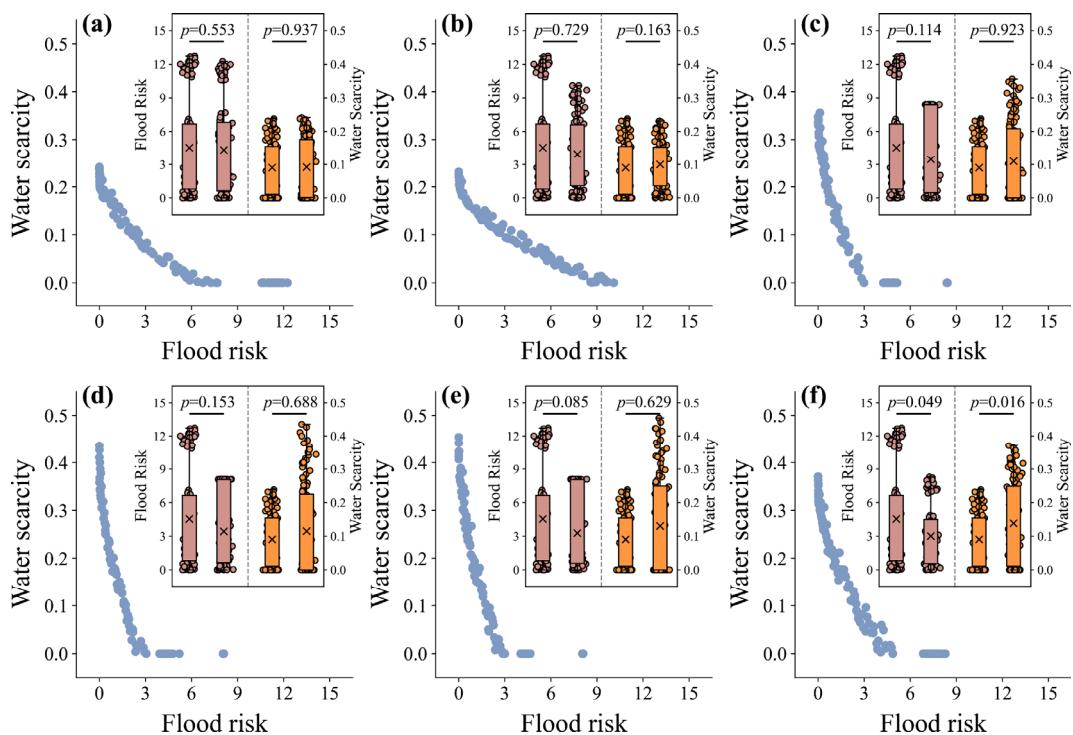
788 objective optimization framework, particularly during the implementation of MAR,
789 how the minimum discharge within the multi-objective optimization framework satisfy
790 both ecological and recharge flow requirements. Consequently, Equation 6 was
791 accordingly modified as follows:

$$792 Q_t \geq Q_t^{eco} + Q_{MAR} \quad (11)$$

793 where Q_{MAR} represents the flow required for different recharge rates $[L^3 T^{-1}]$, while other
794 variables remain consistent with Equation 6. This modification ensures that the
795 minimum release constraint not only maintains downstream ecological flow but also
796 guarantees sufficient water availability for MAR implementation.

797 Following this adjustment, the relationship between flood risk and water scarcity was
798 examined, with results presented in [Figure 12](#). Despite the adjustment in constraints,
799 the inverse relationships between flood risk and water scarcity persisted. In the multi-
800 objective optimization process, MAR was represented as an increased minimum
801 discharge constraint (Equation 11). Under this constraint, the simulations showed that
802 MAR led to lower average flood risk but higher average reservoir water scarcity. This
803 outcome is evident in all the inset box plots of Figure 12, where applying MAR shifts
804 the mean values (cross symbols) downward for flood risk (brown columns, left without
805 MAR to right with MAR) and upward for water scarcity (orange columns). This change
806 might result from the increase in the minimum discharge constraint under MAR, which
807 needs more water to be released downstream, thereby reducing flood risk while
808 simultaneously increasing reservoir water scarcity losses at the end of the flood season.
809 Statistical tests, as indicated by the p -values shown in each inset box plot, further

810 confirmed that these differences were significant ($p < 0.05$) only at the highest recharge
 811 rate of $300 \text{ m}^3 \text{ d}^{-1}$ (Figure 12f), while no significant differences ($p > 0.05$) were
 812 observed at lower recharge rates (Figures 12a–e). This is because at lower recharge
 813 rates, the changes in the minimum discharge constraint were relatively small, resulting
 814 in limited impacts on the optimization outcomes.



815
 816 **Figure 12. The relationship between flood risk and water scarcity under different**
 817 **managed aquifer recharge (MAR) rates. (a)-(f) show the results under recharge**
 818 **rates of $50, 100, 150, 200, 250$, and $300 \text{ m}^3 \text{ d}^{-1}$, respectively. Inset box plots compare**
 819 **flood risk (brown column) and water scarcity (orange column) without (left bar)**
 820 **and with (right bar) recharge measures. Cross (x) represents average value.**
 821 **Statistical tests p -values are also indicated in the inset box.**

822 Figure 12. The relationship between flood risk and water scarcity under different
 823 managed aquifer recharge (MAR) rates. (a)-(f) show the results under recharge rates of
 824 $50, 100, 150, 200, 250$, and $300 \text{ m}^3 \text{ d}^{-1}$, respectively. Inset box plots compare flood risk
 825 (brown column) and water scarcity (orange column) without (left bar) and with (right
 826 bar) recharge measures. Cross (x) represents average value. Statistical tests p -values

827 are also indicated in the inset box. Similar findings have been reported in previous
828 studies. Schäffer et al. (2022) emphasized that maximum discharge constraints
829 significantly affect the marginal water values (i.e., the opportunity cost of storing water
830 for future power generation) as well as the operational strategies of hydropower systems.
831 Helseth et al. (2022) demonstrated that environmental constraints could complicate
832 scheduling problems by introducing state dependencies and non-convexities.
833 Increasing downstream water demand could improve flood utilization efficiency as
834 shown by Wang et al. (2022). These studies highlighted that variations in constraints
835 influence optimization outcomes. Thus, defining and incorporating constraints, such as
836 those introduced by MAR in this study, is crucial for achieving robust and balanced
837 reservoir optimization outcomes, ensuring that floodwater utilization aligns with
838 multiple objectives including ecological sustainability and groundwater recovery.

839 **4.3 Impact of Decision Variable on Optimization Results**

840 To assess the robustness of the optimization scheme, we evaluated the sensitivity of
841 the results to the choice of decision variable. In Section 2.3.3, we initially used the
842 outflow-to-inflow ratio as the decision variable (Decision Variable 1). As Yang et al.
843 (2017) demonstrated, the efficacy of decision variables can differ significantly across
844 scenarios, especially under uncertain or extreme events, potentially leading to
845 suboptimal outcomes. Therefore, we introduced an alternative decision variable, i.e.,
846 the outflow relative to the previous day's reservoir storage (Decision Variable 2).
847 Subsequently, we applied this new decision variable to multi-objective optimization
848 simulations under three FLWL scenarios, which were described in Section 2.3.1. The

849 comparative results are presented in Table 2.

850 All objective values exhibited significant differences ($p < 0.05$) between the two
851 decision variables across the three FLWL scenarios. Notably, minimum flood risk and
852 reservoir water scarcity no longer reached zero in any scenario under Decision Variable
853 2. Despite these quantitative differences, the variation trends of the objectives across
854 scenarios remained consistent with those obtained under Decision Variable 1.
855 Specifically, as the FLWL increased (from Scenario *a* to *c*), the mean values of all
856 objectives decreased. Moreover, the inverse and positive relationships among flood risk,
857 reservoir water scarcity, and groundwater recovery (see Supporting Information Figure
858 S4) followed the same patterns described in Section 3.3. This indicates that while the
859 choice of decision variable can significantly influence the absolute value of the
860 objective functions, it does not alter the intrinsic interactions among the objectives.

861 **Table 2. Comparison of optimization objectives under different decision**
862 **variables. Decision Variable 1 and 2 are the outflow-to-inflow ratio and the**
863 **outflow relative to the previous day's reservoir storage, respectively.**

Objectives	Scenarios	Decision Variable 1		Decision Variable 2		<i>p</i>
		Average	Range	Average	Range	
Flood Risk	<i>a</i>	17.33	[0, 45.65]	12.63	[4.60, 25.56]	<0.05
	<i>b</i>	5.53	[0, 13.29]	8.56	[4.56, 15.33]	<0.05
	<i>c</i>	4.54	[0, 12.73]	7.85	[4.61, 13.07]	<0.05
Water Scarcity	<i>a</i>	0.23	[0, 0.49]	0.36	[0.21, 0.51]	<0.05
	<i>b</i>	0.11	[0, 0.23]	0.28	[0.11, 0.54]	<0.05
	<i>c</i>	0.09	[0, 0.22]	0.23	[0.14, 0.36]	<0.05
Groundwater Recovery	<i>a</i>	6.92	[5.61, 8.42]	9.61	[7.67, 12.16]	<0.05
(10^{-5})	<i>b</i>	6.57	[5.56, 7.26]	7.55	[3.75, 10.29]	<0.05
	<i>c</i>	4.42	[1.06, 7.17]	4.42	[2.55, 5.81]	<0.05

864 **4.4 Advantages, Limitations and Potential Risks**

865 The developed deep learning and multi-objective optimization scheme, leveraging

866 an encoder-decoder LSTM architecture, offers advantages in groundwater level
867 prediction and optimization. Deep learning models excel at capturing complex non-
868 linear dependencies from historical data. This is especially advantageous when
869 complete data on source-sink terms are difficult to obtain, as the utilized encoder-
870 decoder LSTM can capture the underlying relationships in the system through data-
871 driven learning ([Solgi et al., 2021](#)). Moreover, deep learning models are
872 computationally efficient, enabling them to process large datasets efficiently and
873 integrate seamlessly into optimization schemes. This makes them valuable for large-
874 scale optimization tasks, whereas traditional physics-based models might be more
875 computationally intensive ([He et al., 2022](#); [Tripathy and Mishra, 2024](#)).

876 Although this study advances a robust framework for multi-objective optimization
877 of flood control, water storage, and groundwater recovery by integrating deep learning
878 and 3D groundwater numerical modeling, several limitations constrain its scope and
879 applicability, highlighting avenues for future work. First, a key limitation of using the
880 deep learning model is its limited adaptability to replacing traditional physical models
881 across diverse scenarios. Unlike process-based models governed by physical laws, the
882 deep learning models rely solely on statistical patterns. Consequently, it may struggle
883 to generalize to new, unseen conditions, especially in extreme hydrological events
884 caused by climate change ([Acuña et al., 2025](#)). Model performance may degrade when
885 applying to extreme conditions, highlighting the need for caution in using the deep
886 learning model as a substitute for traditional physical models.

887 Second, the framework's exclusive focus on flood-season operations overlooks

888 critical non-flood season dynamics, particularly the regulation of groundwater
889 extraction, which exacerbates depletion in vulnerable depression cone regions. Under
890 climate change, integrating cross-seasonal strategies, such as augmenting flood-season
891 recharge, while curtailing non-flood extraction through adaptive pumping controls
892 ([Balerna et al., 2024](#); [Tang et al., 2024](#)), could yield more sustainable outcomes by
893 balancing annual water budgets and mitigating long-term groundwater stress.

894 Finally, while MAR emerges as a promising tool for floodwater utilization, it
895 introduces potential environmental risks, including water quality degradation and
896 clogging, which could undermine ecological health if not rigorously managed ([Fiori et](#)
897 [al., 2025](#); [Guo et al., 2023](#)). To mitigate these issues, infiltration and pulsed injection
898 can be applied ([Page et al., 2014](#); [Rodríguez et al., 2018](#)), although these measures
899 increase operational costs and limit economic feasibility. This necessitates expanded
900 analyses, explicitly weighing recharge efficiency against water quality through coupled
901 hydrogeochemical modeling.

902 Despite these limitations and potential risks, the proposed framework's broader
903 applicability remains promising, as FU under diverse conditions ([Ding et al., 2023](#); [Liu](#)
904 [et al., 2015](#)) and groundwater recharge practices ([Alam et al., 2020](#); [Zhang et al., 2020](#))
905 are well-established, supporting the effective integration of flood mitigation and
906 groundwater recovery. Therefore, future work can evolve this framework into a more
907 comprehensive, resilient tool for integrated water resource management applicable
908 beyond the North China Plain, particularly under intensifying climate stress.

909 **5. Conclusions**

910 This study advances integrated water resource management by developing a novel
911 coupled framework that merges deep learning (encoder-decoder LSTM) with multi-
912 objective optimization (NSGA-II) and groundwater numerical modeling (MODFLOW)
913 to optimize floodwater utilization in the Lincheng Reservoir system, North China Plain.
914 The framework targets three key objectives: minimizing flood risk (defined as the
915 cumulative exceedance of reservoir water levels above the flood limited water level
916 (FLWL) during the flood season), reducing reservoir water scarcity (measured as the
917 deviation from normal storage levels at the end of the flood season for reservoir), and
918 maximizing groundwater recovery (quantified as the increase in groundwater levels at
919 the end of the flood season resulting from varied reservoir discharge flows during the
920 flood season). By dynamically adjusting reservoir operations across scenarios while
921 ensuring ecological flows and downstream flood safety, the framework demonstrates a
922 pathway for converting flood hazards into resources for groundwater recovery in
923 vulnerable groundwater depression cone areas. The key conclusions are as follows:

924 1. The encoder-decoder LSTM model exhibited high predictive accuracy for
925 groundwater levels, with optimal performance at a 6-day input sequence ($R^2 = 0.99$
926 and 0.93 respectively for training and testing) and robust results at 9 days ($R^2 =$
927 0.97 and 0.92). This highlights the reliability of the framework in capturing
928 temporal hydrological dependencies.

929 2. Across all scenarios, flood risk exhibited inverse relationships with both water
930 scarcity and groundwater recovery. Higher flood risk, driven by retaining more

931 water in the reservoir, reduced reservoir water scarcity through increased storage
932 but limited groundwater recovery by decreasing discharge volumes, thereby
933 reducing the water available for downstream recharge. Increasing the FLWL
934 weakened these inverse relationships, achieving significant reductions in reservoir
935 flood risk and water scarcity (84.9 and 61.9% respectively), with a moderate
936 decrease (22.2%) in groundwater recovery due to lower discharge volumes.

937 3. Maintaining ecological flows enabled groundwater recovery even under reduced
938 total discharges, emphasizing that continuous river connectivity, rather than
939 volume alone, drives groundwater recharge. This finding challenges conventional
940 volume-focused strategies, advocating for flow continuity as an effective way for
941 groundwater recovery.

942 4. The MODFLOW model, with accurate replication of spatiotemporal
943 groundwater variations (R^2 of 0.82–0.96), validated managed aquifer recharge
944 (MAR) as an effective enhancement for groundwater recovery in depression cones.

945 At 300 $\text{m}^3 \text{ d}^{-1}$ operated on 20 recharge wells during the 2023 flood season,
946 maximum recovery reached 0.46 m, with effective recovery ($>0.01 \text{ m}$) in 17.6% of
947 the area. Incorporating MAR modified discharge constraints, resulting in lower
948 flood risks and increased water scarcity, illustrating constraint-driven trade-offs
949 that must be balanced in adaptive management. This quantifies MAR's efficacy but
950 highlights spatial heterogeneity and the need for site-specific optimization to
951 maximize benefits.

952 5. The proposed framework shows potential for broader application beyond the

953 North China Plain. By transforming flood hazards into recoverable groundwater
954 resources through integrated reservoir operations and groundwater recovery
955 measures, the framework offers a promising strategy for regions facing flood and
956 groundwater depletion risks, advancing climate-resilient water management.

957 **Acknowledgments**

958 This work was supported by the National Natural Science Foundation of China (grant
959 numbers: 4247022276, 42077168) and the National Key R&D Program of China (grant
960 numbers: 2023YFC3006503, 2024YFC3211600). We are grateful to Lei Li from the
961 Ninth Geological Brigade of Hebei Bureau of Geology and Mineral Resources, as well
962 as Yinguo Ren and Yuanyuan Li from the Hebei Provincial Hydrologic Survey and
963 Research Center, for their invaluable assistance in data collection.

964 **References:**

965 Acuña E., E., Loritz, R., Kratzert, F., Klotz, D., Gauch, M. Álvarez C., M. et al., 2025.
966 Analyzing the generalization capabilities of a hybrid hydrological model for extrapolation
967 to extreme events. *Hydrology and Earth System Sciences*, 29(5): 1277-1294. DOI:
968 10.5194/hess-29-1277-2025
969 Alam, S., Gebremichael, M., Li, R.P., Dozier, J. and Lettenmaier, D.P., 2020. Can Managed
970 Aquifer Recharge Mitigate the Groundwater Overdraft in California's Central Valley?
971 *Water Resources Research*, 56(8). DOI: 10.1029/2020WR027244
972 Asher, M.J., Croke, B.F.W., Jakeman, A.J. and Peeters, L.J.M., 2015. A review of
973 surrogate models and their application to groundwater modeling. *Water Resources*

974 Research, 51(8): 5957-5973. DOI: 10.1002/2015WR016967

975 Balerna, J.A., Kramer, A.M., Landry, S.M., Rains, M.C. and Lewis, D.B., 2024. Wetland
976 hydrological change and recovery across three decades of shifting groundwater
977 management. *Journal of Hydrology*, 644: 132052. DOI: 10.1016/j.jhydrol.2024.132052

978 Bellu, A., Sanches Fernandes, L.F., Cortes, R.M.V. and Pacheco, F.A.L., 2016. A framework
979 model for the dimensioning and allocation of a detention basin system: The case of a flood-
980 prone mountainous watershed. *Journal of Hydrology*, 533: 567-580. DOI:
981 10.1016/j.jhydrol.2015.12.043

982 Bermúdez, M., Farfán, J.F., Willems, P. and Cea, L., 2021. Assessing the Effects of Climate
983 Change on Compound Flooding in Coastal River Areas. *Water Resources Research*, 57(10).
984 DOI: 10.1029/2020WR029321

985 Blöschl, G., Hall, J., Parajka, J., Perdigão, R.A.P., Merz, B., Arheimer, B., et al., 2017. Changing
986 climate shifts timing of European floods. *Science*, 357(6351): 588-590. DOI:
987 10.1126/science.aan2506

988 Condon, L. E. and Maxwell, R. M., 2019. Simulating the sensitivity of evapotranspiration and
989 streamflow to large-scale groundwater depletion. *Science Advances*, 5(6): eaav4574. DOI:
990 10.1126/sciadv.aav4574

991 Condon, L.E., Kollet, S., Bierkens, M.F.P., Fogg, G.E., Maxwell, R.M., Hill, M.C., Fransen, H.
992 H., Verhoef, A., Van Loon, A., Sulis, M. and Abesser, C., 2021. Global groundwater
993 modeling and monitoring: Opportunities and challenges. *Water Resources Research*,
994 57(12). DOI: 10.1029/2020WR029500

995 Chagas, V.B.P., Chaffé, P.L.B. and Blöschl, G., 2022. Climate and land management accelerate

996 the Brazilian water cycle. *Nature Communications*, 13(5136). DOI: 10.1038/s41467-022-
997 32580-x

998 Chen, X., Wang, P., Muhammad, T., Xu, Z. and Li, Y., 2020. Subsystem-level groundwater
999 footprint assessment in North China Plain – The world's largest groundwater depression
1000 cone. *Ecological Indicators*, 117: 106662. DOI: 10.1016/j.ecolind.2020.106662

1001 Cui, X., Wang, Z., Xu, N., Wu, J. and Yao, Z., 2024. A secondary modal decomposition
1002 ensemble deep learning model for groundwater level prediction using multi-data.
1003 *Environmental Modelling & Software*, 175: 105969. DOI: 10.1016/j.envsoft.2024.105969

1004 Deb, K., Pratap, A., Agarwal, S. and Meyarivan, T., 2002. A fast and elitist multiobjective
1005 genetic algorithm: NSGA-II. *IEEE Transactions On Evolutionary Computation*, 6(2): DOI:
1006 182-197. 10.1109/4235.996017

1007 Deng, Y., Wang, L., Jia, H., Tong, X. and Li, F., 2019. A Sequence-to-Sequence Deep Learning
1008 Architecture Based on Bidirectional GRU for Type Recognition and Time Location of
1009 Combined Power Quality Disturbance. *IEEE Transactions on Industrial Informatics*, 15(8):
1010 4481-4493. DOI: 10.1109/TII.2019.2895054.

1011 Ding, W., Zhang, C., Lin, J., Zhao, X., Ning, Y. and Zhou, H., 2023. Flood Risk Quantification,
1012 Transmission, and Propagation Analysis for Flood Water Utilization of Parallel Reservoirs.
1013 *Journal of Hydrology*, 618: 129202. DOI: 10.1016/j.jhydrol.2023.129202

1014 Fernandes, V.J., de Louw, P.G.B., Bartholomeus, R.P. and Ritsema, C.J., 2024. Machine
1015 learning for faster estimates of groundwater response to artificial aquifer recharge. *Journal*
1016 *of Hydrology*, 637: 131418. DOI: 10.1016/j.jhydrol.2024.131418

1017 Fiori, A., de Barros, F.P.J. and Bellin, A., 2025. An Analytical Framework for Risk Evaluation

1018 and Design of Infiltration Basins for Managed Aquifer Recharge. Water Resources
1019 Research, 61(1): e2024WR038516. DOI: 10.1029/2024WR038516

1020 Gauch, M., Kratzert, F., Klotz, D., Nearing, G., Lin, J. and Hochreiter, S., 2021. Rainfall–runoff
1021 prediction at multiple timescales with a single Long Short-Term Memory network.
1022 Hydrology and Earth System Sciences, 25(4): 2045-2062. DOI: 10.5194/hess-25-2045-
1023 2021

1024 Guo, Z., Fogg, G. E., Chen, K., Pauloo, R. and Zheng, C., 2023. Sustainability of Regional
1025 Groundwater Quality in Response to Managed Aquifer Recharge. Water Resources
1026 Research, 59(1): e2021WR031459. DOI: DOI: 10.1029/2021WR031459

1027 Haaf, E., Giese, M., Reimann, T. and Barthel, R., 2023. Data-Driven Estimation of
1028 Groundwater Level Time-Series at Unmonitored Sites Using Comparative Regional
1029 Analysis. Water Resources Research, 59(7): e2022WR033470. DOI:
1030 10.1029/2022WR033470

1031 Han, Y., Dong, Z., Cui, C., Zhang, T. and Luo, Y., 2025. Multi-objective optimization
1032 scheduling for extensive plain lake water resources incorporating flood resource utilization.
1033 Journal of Hydrology, 651: 132584. DOI: 10.1016/j.jhydrol.2024.132584

1034 He, S., Guo, S., Zhang, J., Liu, Z., Cui, Z. and Zhang, Y. et al., 2022. Multi-objective operation
1035 of cascade reservoirs based on short-term ensemble streamflow prediction. Journal of
1036 Hydrology, 610: 127936. DOI: 10.1016/j.jhydrol.2022.127936

1037 Hirabayashi, Y., Mahendran, R., Koirala, S., Konoshima, L., Yamazaki, D., Watanabe, S., Kim,
1038 H. and Kanae, S., 2013. Global flood risk under climate change. Nature Climate Change,
1039 3(9): 816-821. DOI: 10.1038/nclimate1911

1040 Hosseini, F., Prieto, C. and Alvarez, C., 2024. Hyperparameter optimization of regional
1041 hydrological LSTMs by random search: A case study from Basque Country, Spain. *Journal*
1042 of Hydrology

1043 643. DOI: 10.1016/j.jhydrol.2024.132003

1044 Helseth, A., Mo, B., Hågenvik, H.O. and Schäffer, L.E., 2022. Hydropower Scheduling with
1045 State-Dependent Discharge Constraints: An SDDP Approach. *Journal of Water Resources*
1046 Planning and Management

1047 148(11): 04022061. DOI: 10.1061/(ASCE)WR.1943-
1048 5452.0001609

1049 Irvine, D.J., Singha, K., Kurylyk, B.L., Briggs, M.A., Sebastian, Y., Tait, D.R. and Helton, A.
1050 M., 2024. Groundwater-Surface water interactions research: Past trends and future
1051 directions. *Journal of Hydrology*, 644: 132061. DOI: 10.1016/j.jhydrol.2024.132061

1052 Jasechko, S., Seybold, H., Perrone, D., Fan, Y., Shamsuddoha, M., Taylor, R.G., Fallatah, O.
1053 and Kirchner, J.W., 2024. Rapid groundwater decline and some cases of recovery in
1054 aquifers globally. *Nature*, 625(7996): 715-721. DOI: 10.1038/s41586-023-06879-8

1055 Ji, X., Ding, J., Shen, H.W. and Salas, J.D., 1984. Plotting positions for Pearson type-III
1056 distribution. *Journal of Hydrology*, 74(1): 1-29. DOI: 10.1016/0022-1694(84)90137-9

1057 Jain, S.K., Shilpa, L.S., Rani, D. and Sudheer, K.P., 2023. State-of-the-art review: Operation of
1058 multi-purpose reservoirs during flood season. *Journal of Hydrology*, 618: 129165. DOI:
1059 10.1016/j.jhydrol.2023.129165

1060 Jiang, Z., Liu, P., Ji, C., Zhang, H. and Chen, Y., 2019. Ecological flow considered multi-
1061 objective storage energy operation chart optimization of large-scale mixed reservoirs.
Journal of Hydrology, 577: 123949. DOI: 10.1016/j.jhydrol.2019.123949

1061 Kuang, X.X., Liu, J.G., Scanlon, B.R., Jiao, J.J., Jasechko, S. and Lancia, M. et al., 2024. The

1062 changing nature of groundwater in the global water cycle. *Science*, 383(6686). DOI:
1063 10.1126/science.adf0630

1064 Li, J., Zhang, W. and Yeh, W.W.G., 2021. A Proposed Multi-Objective, Multi-Stage Stochastic
1065 Programming with Recourse Model for Reservoir Management and Operation. *Water
1066 Resources Research*, 57(10): e2020WR029200. DOI: 10.1029/2020WR029200

1067 Liu, P., Li, L., Guo, S., Xiong, L., Zhang, W., Zhang, J. and Xu, C., 2015. Optimal design of
1068 seasonal flood limited water levels and its application for the Three Gorges Reservoir.
1069 *Journal of Hydrology*, 527: 1045-1053. DOI: 10.1016/j.jhydrol.2015.05.055

1070 Lyu, K., Dong, Y., Lyu, W., Zhou, Y., Wang, S., Wang, Z., Cui, W., Zhang, Y., Zhang, Q. and
1071 Cui, Y., 2025. Data-driven and numerical simulation coupling to quantify the impact of
1072 ecological water replenishment on surface water-groundwater interactions. *Journal of
1073 Hydrology*, 649: 132508. DOI: 10.1016/j.jhydrol.2024.132508

1074 Malhotra, P., Ramakrishnan, A., Anand, G., Vig, L., Agarwal, P. and Shroff, G.M., 2016.
1075 LSTM-based Encoder-Decoder for Multi-sensor Anomaly Detection. Arxiv Preprint
1076 Arxiv:1607.00148. DOI:10.48550/arXiv.1607.00148

1077 Mateo, C. M., Hanasaki, N., Komori, D., Tanaka, K., Kiguchi, M., Champathong, A.,
1078 Sukhapunnaphan, T., Yamazaki, D. and Oki, T., 2014. Assessing the impacts of reservoir
1079 operation to floodplain inundation by combining hydrological, reservoir management, and
1080 hydrodynamic models. *Water Resources Research*, 50(9): 7245-7266. DOI:
1081 10.1002/2013WR014845

1082 Mu, Z.Y., Ai, X.S., Ding, J., Huang, K., Chen, S.L., Guo, J.J. and Dong, Z., 2022. Risk Analysis
1083 of Dynamic Water Level Setting of Reservoir in Flood Season Based on Multi-index.

1084 Water Resources Management, 36(9): 3067-3086. DOI: 10.1007/s11269-022-03188-z

1085 Page, D., Vanderzalm, J., Miotliński, K., Barry, K., Dillon, P., Lawrie, K., et al., 2014.

1086 Determining treatment requirements for turbid river water to avoid clogging of aquifer

1087 storage and recovery wells in siliceous alluvium. Water Research, 66:99-110. DOI:

1088 10.1016/j.watres.2014.08.018

1089 Rodríguez, E., P., Canelles, A., Sanchez-Vila, X., Folch, A., Kurtzman, D., Rossetto, R., et al.,

1090 2018. A risk assessment methodology to evaluate the risk failure of managed aquifer

1091 recharge in the Mediterranean Basin. Hydrology and Earth System Sciences, 22(6): 3213-

1092 3227. DOI: 10.5194/hess-22-3213-2018

1093 Owuor, S.O., Butterbach-Bahl, K., Guzha, A.C., Rufino, M.C., Pelster, D.E., Díaz-Pinés, E. and

1094 Breuer, L., 2016. Groundwater recharge rates and surface runoff response to land use and

1095 land cover changes in semi-arid environments. Ecological Processes, 5(1): 16. DOI:

1096 10.1186/s13717-016-0060-6

1097 Roy, D.K., Leslie, D.L., Reba, M.L., Hashem, A.A., Bellis, E. and Nowlin, J., 2024. Optimizing

1098 the quantity of recharge water into a sedimentary aquifer through infiltration galleries

1099 using a surrogate assisted coupled simulation–optimization approach. Journal of

1100 Hydrology, 635: 131183. DOI:10.1016/j.jhydrol.2024.131183

1101 Hochreiter, S. and Schmidhuber, J., 1997. Long Short-Term Memory. Neural Computation, 9(8):

1102 1735-1780. DOI: 10.1162/neco.1997.9.8.1735.

1103 Verma, S., Pant, M. and Snasel, V. 2021. A Comprehensive Review on NSGA-II for Multi-

1104 Objective Combinatorial Optimization Problems. IEEE Access, 9: 57757-57791. DOI:

1105 10.1109/ACCESS.2021.3070634

1106 Samanta, S., Sheng, Z., Munster, C.L. and Van Houtte, E., 2020. Seasonal variation of
1107 infiltration rates through pond bed in a managed aquifer recharge system in St-André,
1108 Belgium. *Hydrological Processes*, 34(18): 3807-3823. DOI: 10.1002/hyp.13827

1109 Schäffer, L.E., Helseth, A. and Korpas, M., 2022. A stochastic dynamic programming model
1110 for hydropower scheduling with state-dependent maximum discharge constraints.
1111 *Renewable Energy*, 194: 571-581. DOI: 10.1016/j.renene.2022.05.106

1112 Solgi, R., Loáiciga, H. A. and Kram, M., 2021. Long short-term memory neural network
1113 (LSTM-NN) for aquifer level time series forecasting using in-situ piezometric
1114 observations. *Journal of Hydrology*, 601: 126800. DOI: 10.1016/j.jhydrol.2021.126800

1115 Sutskever I., Vinyals, O. and Le, Q.V., 2014. Sequence to Sequence Learning with Neural
1116 Networks. Arxiv E-Prints: arXiv:1409.3215. DOI: 10.48550/arXiv.1409.3215Tabari, H.,
1117 2020. Climate change impact on flood and extreme precipitation increases with water
1118 availability. *Scientific Reports*, 10(1): 13768. DOI: 10.1038/s41598-020-70816-2

1119 Tang, W., Zhao, X., Wang, J., Motagh, M., Xu, H., Ru, Z. and Wang, Y., 2024. Land surface
1120 response to groundwater drawdown and recovery in Taiyuan city, Northern China,
1121 analyzed with a long-term elevation change measurements from leveling and multi-sensor
1122 InSAR. *Journal of Hydrology*, 641: 131781. DOI: 10.1016/j.jhydrol.2024.131781

1123 Tennant, D.L., 1976. Instream Flow Regimens for Fish, Wildlife, Recreation and Related
1124 Environmental Resources. *Fisheries*, 1(4): 6-10. DOI: 10.1577/1548-8446(1976)
1125 001<0006: IFRFFW>2.0.CO;2

1126 Tripathy, K.P. and Mishra, A.K., 2024. Deep learning in hydrology and water resources
1127 disciplines: concepts, methods, applications, and research directions. *Journal of Hydrology*,

1128 628: 130458. DOI: 10.1016/j.jhydrol.2023.130458

1129 Wang, K., Wang, Z., Liu, K., Du, H., Mo, Y., Wang, M., Jin, G. and Li, L., 2022. Assessing the
1130 floodwater utilization potential in a reservoir-controlled watershed: A novel method
1131 considering engineering regulations and an empirical case from China. Ecological
1132 Informatics, 68: 101581. DOI: 10.1016/j.ecoinf.2022.101581

1133 Wang, K., Wang, Z.Z., Xu, J., Mo, Y.M., Cheng, L., Bai, Y., Du, H.H. and Liu, K.L., 2023.
1134 Floodwater utilization potential assessment of China based on improved conceptual model
1135 and multi-reservoir basin assessment method. Journal of Hydrology, 625. DOI:
1136 10.1016/j.jhydrol.2023.129921

1137 Wei, G., Liang, G., Ding, W., He, B., Wu, J., Ren, M. and Zhou, H., 2022. Deriving optimal
1138 operating rules for flood control considering pre-release based on forecast information.
1139 Journal of Hydrology, 615: 128665. DOI: 10.1016/j.jhydrol.2022.128665

1140 Wunsch, A., Liesch, T. and Broda, S., 2021. Groundwater level forecasting with artificial neural
1141 networks: a comparison of long short-term memory (LSTM), convolutional neural
1142 networks (CNNs), and non-linear autoregressive networks with exogenous input (NARX).
1143 Hydrology and Earth System Sciences, 25(3): 1671-1687. DOI: 10.5194/hess-25-1671-
1144 2021

1145 Xiang, Z., Yan, J. and Demir, I., 2020. A Rainfall-Runoff Model With LSTM-Based Sequence-
1146 to-Sequence Learning. Water Resources Research, 56(1): e2019WR025326. DOI:
1147 10.1029/2019WR025326

1148 Yang, G., Guo, S., Liu, P., Li, L. and Xu, C., 2017. Multiobjective reservoir operating rules
1149 based on cascade reservoir input variable selection method. Water Resources Research,

1150 53(4): 3446-3463. DOI: 10.1002/2016WR020301

1151 Yang, S., Zhao, B., Yang, D., Wang, T., Yang, Y., Ma, T. and Santisirisomboon, J., 2023. Future
1152 changes in water resources, floods and droughts under the joint impact of climate and land-
1153 use changes in the Chao Phraya basin, Thailand. *Journal of Hydrology*, 620: 129454. DOI:
1154 10.1016/j.jhydrol.2023.129454

1155 Zhang, C., Duan, Q. Y., Yeh, P. J. F., Pan, Y., Gong, H. L., Gong, W., Di, Z., Lei, X., Liao, W.,
1156 Huang, Z., Zheng, L. and Guo, X., 2020. The Effectiveness of the South-to-North Water
1157 Diversion Middle Route Project on Water Delivery and Groundwater Recovery in North
1158 China Plain. *Water Resources Research*, 56(10): e2019WR026759. DOI:
1159 10.1029/2019WR026759

1160 Zhuang, D., Gan, V.J.L., Tekler, Z.D., Chong, A.D., Tian, S. and Shi, X., 2023. Data-driven
1161 predictive control for smart HVAC system in IoT-integrated buildings with time-series
1162 forecasting and reinforcement learning. *Applied Energy*, 338. DOI:
1163 10.1016/j.apenergy.2023.120936